Enhancing DNA metabarcoding performance and applicability with bait capture enrichment and DNA from conservative ethanol

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TITLE: Enhancing DNA metabarcoding performance and applicability with bait capture enrichment and DNA from conservative ethanol

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Capture enrichment & etDNA for metabarcoding

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ABSTRACT

Metabarcoding is often presented as an alternative identification tool to compensate for coarse taxonomic resolution and misidentification encountered with traditional morphological approaches. However, metabarcoding comes with two major impediments which slow down its adoption. First, the picking and destruction of organisms for DNA extraction are time and cost consuming and do not allow organism conservation for further evaluations. Second, current metabarcoding protocols include a PCR enrichment step which induces errors in the estimation of species diversity and relative biomasses. In this study, we first evaluated the capacity of capture enrichment to replace PCR enrichment using

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controlled freshwater macrozoobenthos mock communities. Then, we tested if DNA extracted from the fixative ethanol (etDNA) of the same mock communities can be used as an alternative to DNA extracted from pools of whole organisms (bulk DNA). We show that capture enrichment provides more reliable and accurate representation of species occurrences and relative biomasses in comparison with PCR enrichment for bulk DNA. While etDNA does not permit to estimate relative biomasses, etDNA and bulk DNA provide equivalent species detection rates. Thanks to its robustness to mismatches, capture enrichment is already an efficient alternative to PCR enrichment for metabarcoding and, if coupled to etDNA, is a time-saver option in studies where presence information only is sufficient.

KEYWORDS

bait capture, ethanol DNA, DNA enrichment, metabarcoding, biomass estimation, macrozoobenthos

INTRODUCTION

Reliable and accurate taxa identification is fundamental in biological sciences. Poor taxonomic identification can lead to cascades of error affecting our knowledge and understanding not only in theoretical and fundamental biology, but also in applied fields leading to poor management decisions (Bortolus, 2008). It distorts our ability to infer processes in ecology and evolution, to manage and conserve human-impacted systems and to carry out human health and resource programs (Bortolus, 2008; Leys, Keller, Räsänen, Gattolliat, & Robinson, 2016; Prié, Puillandre, & Bouchet, 2013). Poor taxonomic identification occurs when identification to species level is not possible (coarse taxonomic resolution) or when identification is incorrect (misidentification). For example, Martin, Adamowicz, and Cottenie (2016) investigated macrozoobenthos community distribution in freshwater streams at different taxonomic level (family, genus and species) and found a spatial structure only when the species identification level was reached. In Vietnam, misidentification in *Anopheles* species lead to mismanagement in control program of the Malaria disease as the main target species was a non-vector one (Van Bortel et al., 2001). This kind of difficulties leading to poor taxonomic identification and thus to misinterpretation and erroneous conclusions is often associated to morphological based taxonomy

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(Baird & Hajibabaei, 2012; Creer et al., 2016). Cryptic species, limited expertise, damaged or juvenile specimens and even cost and time constraints in the case of applied research are all problems encountered in morphological identification leading to misidentification or coarse taxonomic resolution (Bringloe, Cottenie, Martin, & Adamowicz, 2016; Hajibabaei, Baird, Fahner, Beiko, & Golding, 2016; Ji et al., 2013; Sweeney et al., 2011).

In the past decade, DNA-based identification has been proposed as an alternative to morphological approaches. DNA barcoding uses the sequence of a genetic marker of one specimen, hereafter called a DNA barcode, usually of an organelle genome for eukaryotes (e.g. mitochondria for animals and chloroplast for plants, see Creer et al. (2016) for overview) and assign it to a species name within a reference database (Hebert, Ratnasingham, & de Waard, 2006). DNA barcoding supposedly addresses limitations in morphological identification by accurately discriminating species regardless of their morphology (Sweeney et al., 2011), development stages (Hubert, Delrieu-Trottin, Irisson, Meyer, & Planes, 2010) or sex (Forshaw, 2010). Recent advances in high throughput technologies enabled the emergence of DNA metabarcoding, the barcoding of pool of specimens in a single reaction which permit to work on whole community at once (Pompanon et al, 2011). Metabarcoding has already been adopted as a routine identification tool in microbial community ecology (Abdelfattah, Malacrinò, Wisniewski, Cacciola, & Schena, 2018). In macro-organism community studies, the last ten years have seen the development of invertebrate derived DNA (iDNA, (Calvignac-Spencer et al., 2013; Schnell et al., 2012)) and the analysis of DNA associated with environmental matrices such as soil or water (eDNA, (Harper et al., 2018; Thomsen & Willerslev, 2015)). Metabarcoding was instrumental in the emergence of these new approaches for which, by definition, there is no non-molecular alternative. Conversely, in conventional macro-organism community studies where the traditional identification using morphological criteria has been used for many years, metabarcoding is still currently marginally used as a routine identification tool. Yet, metabarcoding could enhance the capacity to characterize these communities through iDNA, eDNA or through the extraction of DNA from a homogenate of whole organisms collected together (bulk DNA; (Deiner et al., 2017)). A major limitation to the democratization of metabarcoding is the absence of standardized metabarcoding protocol that has been

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established and validated by the scientific community (Leese et al., 2018). In particular, methodological roadblocks are encountered throughout each step of sample processing (i.e. DNA extraction, enrichment of the targeted DNA barcode, sequencing, bioinformatic treatment and taxonomic assignment). Hereafter, we focused on two major metabarcoding roadblocks: specimen picking before extraction and DNA barcode enrichment.

Specimen picking is a major concern for bulk DNA samples and is particularly critical when the ratio of targeted organism over substrate is low (e.g. invertebrates sampled in streambed). It aims to separate individuals from substrate (e.g. leaves, sand…) prior to DNA extraction to avoid PCR inhibition (Elbrecht, Vamos, Meissner, Aroviita, & Leese, 2017) and to limit the quantity of material to be processed during DNA extraction. This step is time and cost consuming. Direct extraction from fixative agent, usually ethanol, has been proposed as a time-saver alternative to specimen picking (Hajibabaei, Spall, Shokralla, & van Konynenburg, 2012; Zizka, Leese, Peinert, & Geiger, 2018). When fixative agent is used as a DNA template, organisms are not destroyed and are conserved for further taxonomic work or downstream analyses (Leese et al., 2016). Similar techniques have already been employed by the ancient DNA community where the destruction of samples such as museum specimens is to be avoided (Paijmans, Fickel, Courtiol, Hofreiter, & Forster, 2016; van der Valk, Lona Durazo, Dalen, & Guschanski, 2017). Yet, little research on bulk DNA has been conducted on this alternative (but see Hajibabaei et al. (2012); Zizka et al. (2018)) and a rigorous comparison with traditional specimen picking is warranted before it can be used as a standard template of DNA in community studies.

PCR enrichment bias is often considered as the most problematic roadblock in metabarcoding because it may alter species detection and relative abundance recovery of species (Elbrecht & Leese, 2017; Leese et al., 2018; Piñol, Senar, & Symondson, 2018). Prior to sequencing, DNA barcodes are first amplified by PCR using primers that may not have the same number of mismatches with the targeted sequences across taxa (Piñol, Mir, Gomez-Polo, & Agustí, 2015). During a PCR, if DNA barcode sequences from two different species are in equimolar concentration, but that the first species

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presents less mismatches with the primers, this species' sequence will be amplified preferentially. Consequently, amplification efficiency is expected to be non-equal across taxa, leading from under-amplification to no amplification of some taxa in the worst case scenario. PCR bias was demonstrated for fungi (Bellemain et al., 2010), bacteria (Frank et al., 2008), invertebrates (Piñol et al., 2015) and vertebrates (Arif, Khan, Al Sadoon, & Shobrak, 2011). Other factors affect PCR like GC content (Aird et al., 2011) or inhibitors that can remain after DNA extraction but primer bias is commonly presented as the major cause of biases in metabarcoding (e.g. (Elbrecht & Leese, 2015; Piñol et al., 2015; Pinto & Raskin, 2012)). In consequence, a lot of work focused on primer design to decrease PCR biases with, for instance, the use of several primer pairs, degenerated primers or amplification of several DNA barcodes (Drummond et al., 2015; Elbrecht & Leese, 2017; Elbrecht et al., 2016; Gibson et al., 2015; Jusino et al., 2019; Leray & Knowlton, 2017; Zhang, Chain, Abbott, & Cristescu, 2018). These efforts increased the species detection rate but the quantitative bias was not solved completely (Piñol et al., 2018).

Avoiding PCR enrichment will, by definition, solve the PCR bias issue (Porter & Hajibabaei, 2018). Low (Linard, Crampton-Platt, Gillett, Vogler, & Timmermans, 2015) or high (Porter & Hajibabaei, 2018) coverage metagenome sequencing (i.e. sequencing a community DNA without any enrichment) can be used to assemble entire organelle genomes. This approach provides an efficient way to recover species richness and taxa relative biomass, although the proportion of organelle reads is extremely low making metagenome sequencing much more expensive than PCR metabarcoding (Bista et al., 2018; Gómez-Rodríguez, Crampton-Platt, Timmermans, Baselga, & Vogler, 2015; Zhou et al., 2013). Furthermore, only a small part of an organelle genome is usable for taxonomic assignment as reference databases mostly contain DNA barcode sequences (e.g. COI for metazoan, 16S for bacteria, ITS for fungi, (Creer et al., 2016)). Although methods are being developed to reduce organelle genome sequencing cost (Macher, Zizka, Weigand, & Leese, 2017), the construction of exhaustive organelle genome reference databases will be a long-term and expensive process. Another PCR-free alternative is capture enrichment where targeted sequences hybridize to baits and are retrieved by magnetism (Dowle,

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Pochon, C. Banks, Shearer, & Wood, 2016). Contrary to metagenome sequencing, capture enrichment increases the proportion of targeted reads reducing the sequencing cost (Jones & Good, 2016). Baits are long oligonucleotides (more than 60 bp) which are designed from reference sequences. Capture enrichment has already been extensively used in genomics and genetics where it is used to retrieve thousands of loci (e.g. exons capture, (Hodges et al., 2007)) and SNPs arrays (Yang et al., 2009) of one species in a single reaction prior to sequencing. It is also commonly used for ancient DNA to both increase the amount of endogenous DNA and allow loci enrichment where DNA fragmentation tends to complicate PCR reactions (Avila-Arcos et al., 2011; Horn, 2012). Capture enrichment has also recently been used in phylogenetics where the goal is to target thousands of loci in many related species (McCormack et al., 2013; Phuong & Mahardika, 2018). Conversely, capture enrichment is not commonly used in community ecology (but see Dowle, Pochon, J, Shearer, & Wood, (2016); Shokralla et al., (2016)) which opens new challenges as in this new context the diversity of organism is extremely high but the number of targeted loci is rather small. Capture enrichment should be more robust to identify taxa in a community than PCR enrichment because (i) thousands of different baits can be designed and (ii) when sequences are unknown or species are polymorphic, few mismatches between the baits and the targeted sequences should not bias DNA enrichment as they would do with PCR primers (Li, Hofreiter, Straube, Corrigan, & Naylor, 2013; Paijmans et al., 2016; Portik, Smith, & Bi, 2016). For example, one bait designed for a species of the genus *Danio* permitted to detect others *Danio* species in the Amazon basin where the reference species was absent (Mariac et al., 2018). In a single study, capture enrichment was compared to PCR enrichment and was found to detect more taxa but was unable to estimate relative abundances (Dowle et al., 2016). Liu et al. (2016) and Wilcox et al. (2018) have shown that relative abundances can be recovered with capture enrichment if species-specific corrections were to be applied to take into account variation in the number of mitochondria copy number among species (Liu et al., 2016; Wilcox et al., 2018). Such corrections are unfortunately not suitable to the complexity of field samples. Dowle et al. (2016) study was based on natural communities only described at a coarse taxonomic level (family to genus level) and with indirect biomass measurements.

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Thus, the capacity of capture enrichment to describe both community diversity, and relative biomass without species-specific biomass corrections, requires further testing with controlled communities.

In this study, we first investigated the capacity of cytochrome oxidase subunit I (COI) gene capture enrichment to detect taxa and retrieve initial biomass without species specific correction. Second, we evaluated if DNA extracted from ethanol (etDNA) can be used as an alternative to organism picking and homogenization (bulk DNA) by assessing species detection and initial biomass recovery of this alternative template DNA with PCR and capture enrichment. Tests were carried out using two types of freshwater mock communities (MC): (i) low diversity MC (10 species) with variable dry biomass across taxa and (ii) high diversity MC (52 taxa) with homogeneous biomass across taxa from (Elbrecht & Leese, 2015).

MATERIALS & METHODS

1. Mock community design

Freshwater macrozoobenthos specimens for the 10 species mock communities (MC) were sampled from various streams in east of France during May 2017, except for two species, *Gammarus fossarum* and *Chironomus riparius*, which came from Irstea livestock, France (ECOTOX team, RiverLy, France). The 10 species were chosen to represent a wide taxonomic range and because they can be easily identified to the species level by the naked eye (table 1). Two hundred milliliters of ethanol 96% (EtOH) were dispense in bottles and individuals were placed alive in the ethanol. Eight samples with different relative biomass for the 10 species were constructed (Figure 1, table 1). Samples were then stored for 6 months at 4°C until DNA extraction. DNA was extracted from the whole organisms (bulk DNA) and from the preservative EtOH (etDNA).

The 52 taxa mock communities were designed by Elbrecht and Leese (2015). In a nutshell, for ten different MC, a roughly similar dry biomass was collected from 52 taxa identified to the lowest taxonomic level based on morphology, and the homogenized tissue pool was then extracted for each

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189 sample using a salt extraction protocol (Figure 1, see Elbrecht & Leese (2015) for details). Because the
190 biomass was fairly homogeneous among taxa in these MC, we only used these MC to compare the
191 performance of capture and PCR enrichment in predicting species occurrences.
192
193 *2. DNA extraction of the 10 species mock communities*
194 *Bulk DNA.* For each MC, individuals were picked up and sorted by species in petri dishes. They were
195 let to dry overnight and dry biomass for each species was weighted. Mollusc shells and Trichoptera
196 cases were removed prior to the weighing. Individuals were pooled together and the entire community
197 was grounded with a bead mill MM200 (Retsch) during 4 min at 30 Hz. The whole homogenate was
198 extracted (mean ± SD: 245.8 ± 28.6 mg) with FastDNA® Spin Kit for Soil (MP Biomedicals, USA)
199 following manufacturer's protocol. The extracted DNA was purified with Agencourt AMPure XP
200 purification beads (Beckman Coulter, USA) to further remove solvents. A negative extraction control
201 was also made, following the same extraction protocol but without starting material.
202
203 *etDNA.* For each mock community, roughly 50 mL of preservative ethanol was collected. Glycogen
204 and sodium acetate were added to precipitate DNA and samples were placed at -80°C for at least 72
205 hours. After centrifugation, ethanol was removed and after total ethanol evaporation, the dried residue
206 was dissolved in the buffer solution of NucleoSpin Tissue® kit (Macherey-Nagel Gmbh, Germany) and
207 DNA was extracted following manufacturer's instructions. As the amount of DNA was limiting,
208 extractions were repeated twice: a first extraction to compare PCR and capture enrichment methods
209 (extraction I); a second to compare two different primer pairs performance for this supposedly degraded
210 DNA template (extraction II). Four negative extraction controls were also made during extraction II by
211 filling tubes with 96% ethanol which were then extracted following the same protocol.
212
213 *3. COI mock community reference databases*
214 The 10 species reference database was built by sequencing the COI Folmer region of each species of
215 the 10 species mock communities (Folmer, Black, Hoeh, Lutz, & Vrijenhoek, 1994). One specimen per

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species was extracted following the same protocol as for bulk DNA. PCR reactions were performed in a total volume of 25 µL with 1X of PCR standard buffer (including 3 mM $MgCl_2$, Eurobio, France), 0.05 U/µL of EurobioTaq DNA polymerase (Eurobio, France), 0.8 mM of each dNTP (Eurogentec, Belgium), 0.1 mg/mL of BSA (New England BioLabs, USA), 0.4 µM of each primer (LCO1490/HCO2198, Table S1) and 0.5 µL of template DNA. The amplification consisted in an initial denaturation at 95°C for 3 min, followed by 40 cycles of denaturation at 95°C for 20 s, annealing at 51°C for 30 s and extension at 72°C for 45 s, with a final extension at 72°C for 5 min. Purified template DNA was sequenced on both strands with the PCR primers using standard Sanger sequencing (Biofidal, France). All sequences were manually checked for errors and cleaned up with Finch TV software 1.4.0. This protocol did not work for one species (*Heptagenia sulphurea*) and its sequence was downloaded from NCBI to complete the 10 species reference database (GenBank Accession Number HE651395.1).

The 52 taxa reference database was already available for the 52 taxa MC (Elbrecht & Leese, 2017). It includes the haplotypes of all organisms used in their experiment, leading to a reference database of 212 COI sequences. For one taxon (Nematoda), Elbrecht and Leese (2017) could not get any amplification. Since this taxon is missing in Elbrecht and Leese (2017) haplotype reference database, it was not taken into account for downstream analysis.

4. Amplicon sequencing and analysis

Library preparation and sequencing. For the 10 species and 52 taxa MC bulk DNA and etDNA (extraction II), a 421 bp fragment within the COI Folmer region was amplified with the BF2/BR2 primer set with Illumina Nextera tails which is well suited for freshwater macrozoobenthos (Elbrecht & Leese, 2017), Table S1). For etDNA extractions I and II, a shorter region of 178 bp (Fwh1 primer set with Illumina Nextera tails, (Vamos, Elbrecht, & Leese, 2017), Table S1) was also targeted. PCR reactions were performed in triplicates in a total volume of 25 µL with 1X of PCR buffer (including 3mM $MgCl_2$ and 400 µM each dNTP, QIAGEN® Multiplex PCR kit, Qiagen, Germany), 0.5 µM of each primer and 10 ng of template DNA for bulk DNA or 5 µL of template DNA for etDNA. The amplification consisted

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in an initial denaturation at 95°C for 5 min, followed by 25 (bulk DNA) or 35 (etDNA) cycles of denaturation at 95°C for 30 s, annealing at 50°C (bulk DNA) or 52°C (etDNA) for 30 s and extension at 72°C for 2 min, with a final extension at 72°C for 10 min. Triplicate PCR products were pooled and purified with Agencourt AMPure XP purification beads (Beckman Coulter, USA) and quantified using QuantiFluor® dsDNA System (Promega, USA). Ten ng of each purified PCR products were then used in a second PCR to dual index each sample with a unique tag combination (Table S3 for tag combinations details) and to add Illumina adapters. PCR reactions were performed in a total volume of 25 µL with 1X of PCR buffer (BIOAmp® Blend Mix, Biofidal, France), 2.5 nM of $MgCL_2$, 200 µM of each dNTP, 0.25 µM of Illumina primer (Table S1) and 0.02 U/µL of HOT BioAmp Taq (Biofidal, France). PCR products were purified with Agencourt AMPure XP purification beads (Beckman Coulter, USA), quantified using QuantiFluor® dsDNA System (Promega, USA) and pooled at the same concentration (2 nM) for sequencing. The PCR amplicon libraries were sequenced using a 2*250 paired-end V3 MiSeq sequencing kit (Biofidal, France). Negative controls were included to every PCR during the library preparation.

Bioinformatic analysis. Reads were delivered demultiplexed and adapter trimmed. The reads of the 52 taxa mock communities, bulk DNA and etDNA of the 10 species mock communities were processed independently with the same bioinformatic pipeline. First, forward and reverse reads were merged with Vsearch 2.8.4 with a minimum overlapping of 10 nucleotides and a maximum difference of the overlapping region of 5 nucleotides (Rognes, Quince, Nichols, Flouri, & Mahé, 2016). Then, the primer regions were removed from the merged reads with cutadapt 1.9.1 (M. Martin, 2014) and the reads were quality filtered (maximum expected error (ee) of 1, minimum length of 200, no Ns allowed) and dereplicated with Vsearch. Sequences observed less than twice (< 2 reads) were removed. Then, chimeras were *de novo* detected and removed with Vsearch. Finally, sequences were clustered with a similarity cutoff of 97% identity into MOTUs (Molecular Operational Taxonomic Units). MOTUs were assigned to species using the 10 species reference database or the 52 taxa reference database with the blastn algorithm (Camacho et al., 2009). Only alignments with an e-value under 1E-10, a query cover

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over 200 bases for bulk DNA and over 90 bases for etDNA and an identity over 97% were conserved as good alignments for further analysis. MOTUs were also compared to a complete COI protein reference database made from 8 taxa (*Asellus aquaticus* ADA69754.1, *Daphnia pulex* AAD33231.1, *Dinocras cephalotes* AGZ03516.1, *Gammarus fossarum* YP_009379680.1, *Physella acuta* YP_008994230.1, *Radix balthica* HQ330989.1, *Sericostoma personatum* AJR19241.1, *Thremma gallicum* AJR19254.1) present in one or both MCs using diamond 0.9.22 (blastx, more sensitive option, e-value threshold of 1E-10 (Buchfink, Xie, & Huson, 2014)). This allowed us to detect COI sequences even if they did not belong to the species used in the MC. We also estimated the rate of contamination by comparing the quality filtered reads to a protein database containing the proteomes of 100 eukaryotic species from Ensembl (Zerbino et al., 2017) and Ensembl Metazoa (Kersey et al., 2017) as well as the proteomes of 837 prokaryotic species retrieved from the Microbial Genome Database for Comparative Analysis (Uchiyama, Mihara, Nishide, & Chiba, 2015) selecting one species per genus. Sequences were assigned to coarse taxonomic groups (archaea, eubacteria, fungi, plant, protist, protostomia− here Arthropoda, Annelida, Brachiopoda, Mollusca, Nematoda and Platyhelminthes − and other metazoa − corresponding to Deuterostomia) using the best diamond hit (blastx, more sensitive option, e-value threshold of 1E-10).

5. Capture sequencing and analysis

Bait design. While one of the primary goals of the present study is to assess capture versus PCR enrichment efficiency using mock communities, we chose to develop a larger set of baits that could be used in future assessment of French freshwater macrozoobenthos diversity. Up to now and aside from non-Chironomidae dipteran, 3,245 macrozoobenthos species belonging to 22 orders have been identified in small and medium streams of France (Aukema & Rieger, 2013; Coppa, 2019; D'Hondt & Ben Ahmed, 2009; Dusoulier, 2008; Gargominy et al, 2011; Grand & Boudot, 2007; Henry & Magniez, 1983; Le Doaré & Vinçon, 2019; Pattée & Gourbault, 1981; Piscart & Bollache, 2012; Queney, 2011; Serra et al, 2015; Souty-Grosset et al, 2006; Thomas, 2019; Tillier, 2019; Vallenduuk, 2004). One hundred twenty two species of six orders can be considered as marginal in streams and were thus

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297 discarded. All the available COI sequences were downloaded for the 3,123 remaining species from
298 GenBank and BOLD with PrimerMiner 0.18 in December 2016 (Elbrecht & Leese, 2016). Four orders
299 had less than half of their known taxa with available COI barcode and were discarded. Within the 12
300 remaining orders, the species without COI sequences were barcoded whenever possible, i.e. when
301 tissues could be obtained. Organisms were extracted using Chelex (BioRad, USA) and the PCR and
302 sequencing conditions were identical to those of section 3 of the methods. Sequences were deposited on
303 GenBank (97 species, GenBank Accession Numbers MK584300:MK584515). At the end, 1,525 species
304 out of 1,689 known species had a DNA barcode available, representing more than 90% of the targeted
305 species (Table S2). For bait development, sequences were processed by order. First, COI sequences
306 were aligned with blastx 2.7.1 to a reference *Drosophila yakubai* COI sequence (Accession Number:
307 NC_001322.1) and identical sequences were collapsed using a perl script
308 (http://github.com/TristanLefebure/collapse_to_uniq_seq). Then, 120 bp baits were constructed *in silico*
309 using BaitFisher with a tilling of 60 bp and a cluster threshold of 5%, leading to a total of 15,038 baits
310 generated from 100,367 unique sequences (Mayer et al., 2016). The COI *in silico* baits were then sent
311 for RNA bait synthesis to Arbor Biosciences (USA).
312
313 *Library preparation, hybridization and sequencing.* Starting quantity of DNA was 1 µg for bulk
314 DNA (10 and 52 taxa MC) and between 14 and 57 ng for etDNA (extraction I). DNA of each sample
315 was sheared into approximately 600 bp nucleotide fragments by ultrasound sonication with a Qsonica
316 Q800R (Qsonica, USA). Library preparation was conducted using NEBNext® Ultra™ II DNA Library
317 Prep Kit for Illumina® (New England BioLabs, USA) following manufacturer's instructions. Briefly,
318 after sheared DNA end repair, the 5' ends were phosphorylated and the 3' ends were A-tailed. Then,
319 Illumina Nextera tails (Table S1) were ligated to the DNA fragments followed by a clean-up and a size
320 selection of 500-700 nucleotides long fragments with Agencourt AMPure XP purification beads
321 (Beckman Coulter, USA). Finally, DNA fragments were amplified to dual index the libraries (Table S3
322 for tag combinations details) and to add Illumina adapters (Table S1). COI capture enrichment was
323 conducted using myBaits® Custom kit following manufacturer's instructions (Arbor Biosciences, USA).

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One hundred ng of library DNA was used for capture enrichment for bulk DNA and 230 ng for etDNA following manufacturer's instructions for degraded DNA. Baits were diluted 10 times and hybridization lasted 24h for both DNA templates with the exception that hybridization for etDNA was done at 55°C instead of 65°C following manufacturer's instructions for degraded DNA. The final library amplification was performed in a total volume of 50 µL per reaction with the KAPA HiFi DNA Polymerase (Kapa Biosystems, USA) with P5 and P7 Illumina primers (Table S1) using the following conditions: initial denaturation at 98°C for 2 min, followed by 21 cycles of denaturation at 98°C for 20 s, annealing at 60°C for 30 s and extension at 72°C for 1 min, with a final extension at 72°C for 5 min. Capture library concentrations were determined by qPCR with a KAPA qPCR kit (KAPA Library Quant Kit, Kapa Biosystems, USA) and pooled at the same concentration for sequencing. The capture libraries were sequenced using a 2*300 paired-end V3 MiSeq sequencing kit (Biofidal, France).

Bioinformatic analysis. Reads were delivered demultiplexed and adapter trimmed. As for PCR analysis, the reads of the 52 taxa MC, bulk DNA and etDNA of the 10 species MC were processed independently with the same bioinformatic pipeline. First, forward and reverse reads were merged with Vsearch with a minimum overlapping of 10 nucleotides and a maximum difference of the overlapping region of 5 nucleotides (Rognes et al., 2016). Because shearing could lead to fragment longer than 600 nucleotides, merged and non-merged sequences were conserved for downstream bioinformatic steps. Reads were then quality filtered (maximum ee of 1, minimum length of 150, 50 Ns allowed) with Vsearch. Because the bait set contained baits for other genes for other projects (i.e. 16S, NAD1, NAD4, NAD5, CYTB, and ATP6) and because their presence in the bait set can alter downstream results, the reads corresponding to these genes were removed from the quality filtered reads in all downstream analysis. They were recovered with a blastn on a reference database (blastn-short, e-value threshold of 1E-10) containing the complete sequence of NAD1, NAD4, NAD5, CYTB and ATP6 genes of 8 species (*Asellus aquaticus* ADA69754.1, *Daphnia pulex* AAD33231.1, *Dinocras cephalotes* AGZ03516.1, *Gammarus fossarum* YP_009379680.1, *Physella acuta* YP_008994230.1, *Radix balthica* HQ330989.1, *Sericostoma personatum* AJR19241.1, *Thremma gallicum* AJR19254.1), the 16S of the 10 species MC

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(Table S1) and the 16S corresponding to the 52 taxa MC (downloaded from NCBI). The remaining reads were assigned to species using the 10 species reference database or the 52 taxa reference database using BLAST algorithm (Camacho et al., 2009). Only alignments with an e-value below 1E-10, a query cover over 250 and an identity over 97 were used for taxonomic assignment. As for amplicon, filtered reads were also compared to a complete COI protein reference database to estimate the total number of COI reads and to a protein database to estimate the rate and origin of contaminations.

6. Capture efficacy

Capture efficacy (sensus (Cha & Thilly, 1993)) was evaluated by measuring the percentage of targeted reads for each capture library (i.e. capture specificity) and the X-fold enrichment (i.e. capture efficiency) as in Maggia et al. (2017). The percentage of targeted reads is the ratio of the number of reads assigned with diamond on the COI protein database or with blastn to a 16S, NAD1, NAD4, NAD5, CYTB and ATP6 nucleotide reference database to the number of quality filtered reads. To estimate the X-fold enrichment, four samples from the 10 species MC were sequenced without any enrichment. The X-fold enrichment was calculated using the ratio of the percentage of targeted reads from the capture library to the percentage of targeted reads from the enrichment-free library.

7. Species detection and initial biomass recovery

To compensate for sequencing effort as well as total biomass variation among samples, read count assigned to species and biomass (mg) were transformed in proportion of reads (i.e. ratio of the number of COI reads assigned to the species to the total number of COI reads assigned to the COI mock community reference database) and proportion of biomass (i.e. ratio of the biomass of the species to the total biomass of the MC), respectively.

Species detection. A species was considered present in a sample when represented by at least one read for the capture sequencing pipeline and one MOTU for the amplicon sequencing pipeline. We

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calculated the detection sensitivity (*S*) which measures the number of detection success to the total number of trials ($S = \frac{totalnumberofspeciesdetected}{numberofsamples*numberoftaxa}$).

Biomass recovery. For the bulk DNA and etDNA of the 10 species MC, logistic models were built to investigate the relationship between read proportion and biomass proportion. For each DNA template and enrichment method, we compared sets of mixed effects models (i.e. no fixed effect and biomass as fixed effect with no random effect, random intercept, random intercept and slope by species or samples) using Aikake's Information Criterion (AIC, (Burnham & Anderson, 2002)). We also summarized to what extent each method was able to predict species biomass by comparing the observed read proportions to the expected read proportions where read proportion perfectly predict biomass proportion (i.e. a y=x relationship). To this aim, we calculated the mean absolute error (MAE) for each method as follow: $\frac{1}{n}\sum|observed - expected|$ (Willmott & Matsuura, 2005). The lower the MAE is, the closer the observed read proportions are to the expected read proportions.

All statistical analyses were conducted with R (R Core team, 2018). All logistic models relating biomass proportion to read proportion were fitted using lme4 package (glmer functions, with a Logit link and a Binomial family, (Bates, Mächler, Bolker, & Walker, 2014)).

RESULTS

1. Sequencing results

For bulk DNA, amplicon sequencing produced 60,337 to 365,331 raw reads and capture sequencing 165,089 to 857,040 raw reads per sample. For etDNA, amplicon sequencing produced 209,425 to 365,331 raw reads for Fwh1 primer pair of extraction I, 219,124 to 314,441 raw reads for Fwh1 primer pair of extraction II, 172,036 to 280,065 raw reads for BF2/BR2 primer pair of extraction II and capture sequencing 217,240 to 911,484 raw reads (Table 2). In both approaches, the capture sequencing effort

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was higher but a higher proportion of reads was discarded through the bioinformatic pipeline, particularly for etDNA (Table 2).

2. Capture efficacy

The percentage of targeted reads after capture enrichment (i.e. capture specificity) ranged from 16.24 to 86.66% (mean ± se: 63.82% ± 24.83) for bulk DNA and from 19.52 to 40.79% (mean ± se: 27.18 ± 8.39) for etDNA (Table 2, Table S5). For bulk DNA, the 52 taxa MC showed higher and more homogeneous capture specificity (mean ± se: 80.21 ± 5.09) in comparison with bulk DNA of the 10 species MC (mean 43.34 ± 24.52). The average X-fold enrichment (i.e. capture efficiency) was 974 (range: 214-2470) meaning that on average 974-fold more targeted reads were sequenced with capture enrichment than without any enrichment.

The percentage of COI reads that were assigned to a MC species was very heterogeneous among DNA template and enrichment methods. Amplicons on bulk DNA gave the best results (mean: 96.45% COI assignment), followed by capture on bulk (62.33 %), amplicon on etDNA (46.36 %) and finally capture on etDNA (33.71 %). Reads that do not match to the reference MC COI database can have multiple origins including contamination from other organisms. The majority of reads were assigned to protostomians for all but one experiment − capture enrichment on etDNA − where half of the reads belonged to eubacteria and the other half to protostomians (Table 2, Figure S2).

3. PCR and capture enrichment for bulk DNA

Species detection. Capture enrichment detected systematically more species (S=0.96) than PCR (S=0.68) among the 10 species MC (Figure 2). In PCR libraries, three species belonging to Gastropoda (*Ancylus fluviatilis* and *Physella acuta*) and Amphipoda *(Gammarus fossarum)* were never detected. With capture enrichment, every species was detected in 5 out of 8 samples (Figure 2). In the 52 taxa MC, we found the same pattern, with capture enrichment (S=0.96) detecting more species than PCR enrichment (S=0.93) (Figure 2). Each enrichment method eventually failed to detect a small but different set of taxa.

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Biomass recovery. Whatever the enrichment method, following the AIC, the best mixed model to predict initial species biomass using species read proportion was the model combining biomass as a fixed effect and species as a random effect on the intercept and slope (Table 3). This suggests that the mean read proportion between species is different independently of their biomass and that the relationship between biomass proportion and read proportion also differs among species (Figure 3). Using this type of mixed models, a significant relationship between biomass and read proportion was found for both enrichment methods (PCR: p-value=0.024; capture: p-value=0.001, Figure 3). The mean absolute errors between observed and expected read proportion if biomass proportions were to be perfectly translated into read proportions (i.e. y=x relationship) were higher for PCR (MAE=0.11) than for capture (MAE=0.056). Under a scenario where there is no relationship between biomass and read proportions and where each species contribute to the same read proportion independently of its biomass (i.e. 1/10 read proportion), the MAE would be of 0.07, again highlighting that the PCR enrichment step wiped out most of the biomass signal. While absolute biomass variations may be lost, we also tested if biomass ranks could be recovered using Spearman's rank correlation coefficient. Again PCR enrichment performed poorly compared to capture enrichment (average Spearman rho 0.53 and 0.67 for PCR and capture enrichment, respectively).

4. *etDNA performance*

As for bulk DNA, capture enrichment detected more species (*S*=0.97) than PCR (*S*=0.85) (Figure 4, extraction I) on the etDNA template. Concerning PCR, the primer pair Fwh1 performs better (S=0.86) than the primer pair BF2/BR2 (S=0.7) on etDNA (Figure 4, extraction II). This was expected as the long BF2/BR2 fragment may be difficult to amplify on degraded DNA. Nevertheless the two primers pairs also failed to detect different species probably indicating a primer bias rather than a degradation problem. To confirm this, we tested if the sample composition was mostly driven by DNA template and enrichment method using a Principal Coordinate Analysis (PCoA). Samples clustered by enrichment method and PCR primers but not by DNA template (Figure 5). Therefore, in this experiment, the DNA

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template had little to no impact compared to the enrichment method and primers. Concerning initial biomass recovery, the same models as for bulk DNA were selected for etDNA using the AIC (Table 3). However, contrary to bulk DNA, for both enrichment methods, no significant relationship between biomass proportion and read proportion was found (PCR: p-value=0.37; capture: p-value=0.58). The difficulty to infer biomass from read proportion using etDNA was also supported by high MAE values (0.102 and 0.104 for PCR and capture enrichment, respectively) and low rank correlations (average Spearman rho= 0.40 and 0.28 for PCR and capture enrichment, respectively).

DISCUSSION

Capture versus PCR enrichment for metabarcoding

Capture enrichment consistently led to higher species detection rates compared to PCR enrichment whatever the DNA template or sample taxonomic complexity. Using capture in the 10 species MC, only one species was missed in one third of the samples where it was at low abundance (dry biomass < 1 mg). In the 52 taxa MC, capture enrichment also globally performed better that PCR enrichment in terms of species detection (20 against 35 failed detections, respectively). Both methods also missed different sets of taxa. In PCR enrichment, the commonly missed taxa systematically presented 1 or 2 mismatches with the primer pair (Figure S1) reinforcing the view that primer mismatches are the main cause for species non-detection (Elbrecht et al., 2017). Regarding capture enrichment, 8 taxa had no bait designed for them (i.e. Ceratopogonidae, Blephariceridae, Dicranota, Simuliidae, Tipulidae, Trombidiformes, Dugesia and *Daphnia pulex*) but only two taxa (*Daphnia pulex* and Dugesia) were almost systematically missed. These two taxa have COI haplotypes that are at a divergence of at least 15% to any bait whereas the 6 detected taxa have haplotypes that are closer to baits originally designed for other taxa (Figure S1). The divergence to the oligonucleotide has already been described has an important parameter for species detection (Liu et al, 2015; Portik et al, 2016; Vallender et al, 2011; van der Valk et al, 2017). In particular, Liu et al (2015) using a pool of 49 taxa found a lower enrichment efficiency when the divergence to the bait was higher than 20%. Portik et al (2016), this time working on a multi-loci capture data set, found a negative linear relationship between the probability of

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483 sequencing a locus and the divergence to the bait. This translated into a sharp decline from 60% to 20%
484 chance of sequencing when the divergence increased from 10% to 20%. Altogether, for both enrichment
485 methods, the divergence to the oligonucleotide (i.e. primer or bait) appears to be a determinant factor
486 for species detection. Nevertheless, capture enrichment is much more robust to this divergence issue as
487 it combines three characteristics that PCR enrichment misses. First, capture enrichment is less sensitive
488 to mismatches (18 mismatches or 15% divergence) than PCR (2 mismatches or 5% divergence). Second,
489 because thousands of different baits are designed for a given 120-base region, the probability to
490 encounter mismatches is reduced with capture. Finally, several 120-base regions (5 in this study leading
491 to a total cover of 600 bases) are targeted, increasing the probability to design at least one functional
492 bait for a given species even if this species was not in the database used to design the baits. Alternatively
493 species detection can be improved in traditional PCR-based metabarcoding by targeting multiple DNA
494 fragment (e.g. (Drummond et al., 2015; Zhang, Chain, Abbott, & Cristescu, 2018)). This is also likely
495 to be true but easier to implement with capture enrichment where baits for different loci can be used
496 simultaneously (e.g. Liu et al, 2016). The ability to detect more robustly known and unknown
497 biodiversity using a single enrichment step is a pivotal step towards a better understanding of ecosystem
498 function and structure.
499
500 Capture enrichment also yields read proportions that are better predictors of species relative
501 biomasses than PCR enrichment for bulk DNA. Both enrichment methods presented a positive
502 relationship between relative biomass proportion and read proportion but this relationship was closer to
503 a linear y=x relationship for capture enrichment. That said, a random species effect still remained in the
504 capture enrichment model. Its origin could be attributed to a bait bias similar to the primer bias
505 encountered in PCR enrichment but it can also be explained by other factors. Indeed, using a
506 mitochondrial marker to reconstruct an entire dry biomass community is eventually doomed to fail.
507 First, the amount of mitochondrial DNA belonging to a species in a sample will vary as a function of
508 the average number of mitochondrial genomes per cell which is likely to vary extensively among
509 species, life stages and tissues (Rooney et al, 2015). Wilcox et al. (2018) found that total initial species

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DNA abundances could be recovered using capture enrichment if a correction was applied to normalize for variation in the number of mitochondrial genome per cell. Second, dry biomass proportions are a very rough estimator of cell numbers. Thus, variability in cell number per biomass and in mitochondria per cell combines to blur the biomass/read proportion relationship. A similar mock community experiment but with precise knowledge of the number of mitochondrial genome per species in a sample is needed to better test for the existence of bait biases. One can argue that even if capture enrichment yields good estimates of each species mitochondrial genome proportion in a sample, this estimate will still be a poor biomass or abundance proxy. Interestingly, here we found that read proportions from gene capture enrichment already provides biomass rank estimates which in many ecological contexts, for instance in bioassessment (Elbrecht et al., 2017), would already be informative and may be sufficient.

Ethanol DNA as a fast alternative to bulk DNA for species detection

In agreement with Martins et al. (2019), our results demonstrate the potential of etDNA to replace bulk DNA for macrozoobenthos samples if only species occurrence data is required. Species detection was not determined by DNA template but by the enrichment method and the primer pair used. Hence, species detection with etDNA was equivalent to species detection with bulk DNA. Nonetheless, no relationship between read proportion and biomass proportion was highlighted for etDNA. For biomass recovery to work with etDNA, each species should release a quantity of DNA in the ethanol proportionally to its biomass. However, the release of DNA in the ethanol might differ among development stages or species. When put alive in ethanol, some individuals regurgitate and release a lot of their DNA (Anderson et al., 2013). For example, in the class Gastropoda, one sub-class has an operculum which isolates them from the environment. When put in ethanol, these gastropods close their shell and the quantity of DNA released will be lower than for other gastropods without operculum. Also, after ethanol fixation, depending on the presence of a shell or a thick exoskeleton, the amount of released DNA during the soaking period may also be extremely variable across taxa. In conclusion, while DNA from ethanol offers a fast and non-destructive way to identify the species in a sample, differences in the way species release their DNA in ethanol seem to prevent the use of ethanol for quantitative estimates.

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Compared to bulk DNA, DNA extracted from preservative ethanol is more fragmented and much less concentrated. In addition to environmental contaminants such as the bacterial contaminants observed in this study, this template may also be more susceptible to reagent contamination and cross-contamination. The adoption of this alternative template may therefore need stricter sampling, sample handling and DNA extraction protocols than the one regularly used in community ecology laboratories.

Optimising capture enrichment for metabarcoding

Despite capture enrichment already delivers better results than PCR enrichment in terms of species detection or biomass prediction, the efficiency and specificity of this emerging approach can still be optimized for metabarcoding purposes. Although capture enrichment was efficient to enrich the COI sequences of the macrozoobenthos present in our samples with an enrichment of at least 214-fold compared to non-enriched libraries, the percentage of targeted reads was highly heterogeneous and low for some samples. The specificity of the capture enrichment was particularly low when the diversity of the sample was low (43% of targeted reads for the 10 species MC compared to 80% for the 52 taxa MC). How the taxonomic diversity of a sample influences the efficiency of capture enrichment remains unexplained and warrants further experiments where the impact of a gradually increasing sample diversity is tested while controlling for other factors such as the extraction protocol or the overall phylogenetic diversity.

The specificity of capture enrichment estimated by the number of COI reads that matched a taxa included in the MC was systematically lower than with PCR enrichment (about 40% of the COI reads did not match a MC taxa). When looking at the coarse taxonomic composition of the COI reads sequenced after capture enrichment, we found that for bulk DNA, most of the reads (> 95%) were assigned to protostomians ruling out the possibility that bacterial or plant COI contaminations alone are reducing capture specificity. DNA from other protostomians that interact with the taxa included in the MC (e.g. preys or parasites) may contaminate the samples and reduce capture specificity. Nevertheless, we would expect the same to happen with PCR enrichment which is not the case. Alternatively, Li, Schroeder, Ko, and Stoneking (2012) found that the robustness of capture to mismatches is also a

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564 drawback when targeting mitochondrial genes: in addition to capture the targeted mitochondrial gene,
565 it also captures nuclear copies of this gene (NUMTs). The number of NUMT loci per genome is highly
566 variable across taxa but can easily reach hundreds to thousands of loci (Hazkani-Covo, Zeller, & Martin,
567 2010). In some species there might therefore be more NUMTs in a cell than there are mitochondrial
568 genomes. Given that baits can capture divergent loci and that there is many different baits, NUMTs may
569 represent a significant burden in mitochondrial capture enrichment. Finally, capture enrichment also
570 increase the sequencing coverage of the flanking regions of the targeted sequences (Portik et al, 2016).
571 With the pipeline used in this study, the reads containing more than 41% of a flanking region will not
572 be classified as a read coming from a targeted region. The lower specificity of capture enrichment
573 observed in our samples, but not observed with PCR enrichment, might therefore be explained by a
574 combined effect of all these factors.

575 For capture enrichment on etDNA, we estimated that almost half of the reads belonged to eubacteria
576 and the other half to protostomians. This is likely to be linked to the regurgitation of gastric microbiota
577 and the release of epidermic microbiota by some species. By this process, bacterial DNA may actually
578 dominate the ethanol DNA. In this experiment, a lower hybridization temperature was used for etDNA
579 (55°C) than for bulk DNA (65°C) in the hope of capturing more fragmented DNA. Retrospectively, this
580 lower temperature is likely to have decreased the specificity of the capture enrichment and increased the
581 representation of bacterial DNA. Indeed, the specificity of capture enrichment is higher for degraded
582 DNA and samples with a lot of contaminants when a higher hybridization temperature is used (Paijmans
583 et al., 2016). Noteworthy, 95% of the bacterial reads did not share any similarity with the mitochondrial
584 COI, again reinforcing the idea that the large amount of bacterial reads in etDNA samples is probably
585 the result of the combination of non-stringent hybridization conditions and bacterial DNA dominance
586 in etDNA.

587 As found in this study and others (e.g. Liu et al., 2016; Mariac et al., 2018; Portik et al, 2015), the
588 divergence between the targeted DNA and the baits is a critical factor in capture enrichment. As such,
589 to recover all taxa in a community, the bait design must be optimized to reduce this divergence. The
590 priority is thus to design baits using a reference database that is exhaustive. While reference DNA

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barcodes are available for most species for some groups, this is far from being the case for many groups (e.g. (Sonet et al., 2013)). Knowing the divergence threshold of 15% (this study) over which species are not captured and detected, alternative strategies could be deployed to barcode key missing taxa in the reference databases. Indeed, a very limited set of baits is sufficient to represent a whole family or a genus if the intra-group genetic diversity is lower than the divergence threshold (Mariac et al., 2018). This diversity is heterogeneous between taxa group, for example the maximum intra-family COI divergence in Gammaridea (Amphipoda) is 30% but is only 17% in Chloroperlidae (Plecoptera). Therefore, prior investigations are needed to establish the diversity of each group and the number of taxa that have to be described before a robust set of baits can be designed. Another alternative would be to design baits from a set of representative COI sequences and mutate them according to a given divergence to obtain a set of baits that can hybridize to most sequences. Such baits would permit to capture non-barcoded or even new species but could also reduce capture specificity by capturing untargeted sequences such as bacterial genes or NUMTs.

In addition to the bait design strategy, other capture enrichment parameters may need further optimization for a metabarcoding application. The hybridization temperature is known to be a major factor controlling capture efficacy (Li et al., 2013). However, Paijmans et al. (2016) compared different hybridization temperatures but unexpectedly did not find any interaction between hybridization temperature and distance to the bait. But Paijmans et al. (2016) used a pool of 21 feline species, with all the probes designed using a single reference species. Therefore their results probably need to be tested on a data-set that is more similar to a metabarcoding data-set, with more divergent taxa and probes designed from a large database, before the interaction between hybridization temperature and distance to the bait can be completely dismissed. The number of pooled libraries in a single hybridization reaction is another potentially important parameter. While pooling many samples in a single capture reaction significantly cuts the hybridization budget, Portik et al. (2016) showed that it significantly reduces the complexity of the sequencing libraries. In a metabarcoding context, this may result in a lower species discovery rate in disfavor of the rare species. In the same vein, the number of capture rounds is a substantial parameter increasing capture efficacy but also the hybridization budget (Li et al., 2013;

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618 Mariac et al., 2018; Templeton et al., 2013; van der Valk et al., 2017). With a second round of capture,
619 Mariac et al. (2018) increased capture specificity from 0.57% to 70.5% and capture efficiency by 40
620 times but with the penalty of doubling hybridization costs. After sequencing capture enriched
621 metabarcoding libraries, another challenge is the bioinformatic treatment of the reads. Indeed,
622 processing capture data targeting one or few DNA barcodes for many species is considerably different
623 than assembling multiple targeted genes for a single organism. To our knowledge, while there is several
624 assembly pipelines that were developed for the latter case (co-assembly of different loci from a single
625 species, (e.g. (Allio et al., 2019)), there is no available bioinformatic pipelines for the former.
626 Assembling metabarcoding capture reads is challenging as it requires the independent assembly of
627 orthologous sequences belonging to different species which may be more or less divergent and more or
628 less polymorphic. An alternative to assembly is to directly map the capture reads on a reference database
629 as done in this study. While functional, this solution is sub-optimal because capture reads do not start
630 and end at the same position as in amplicon sequencing. This complicates the read clustering and
631 taxonomic assignment, with many capture reads probably lost as they only partially map to the reference
632 library. While the sequencing of the flanking regions of targeted region would probably be highly
633 valuable for taxonomic assignment, as long as there is no dedicated bioinformatic solution, it
634 paradoxically complicates the processing and analysis of capture reads. Finally, a comparison of these
635 mock community results with field samples in which the diversity is much higher and where the template
636 is also much more complex is warranted to confirm the potential of capture enrichment in community
637 ecology.
638
639 CONCLUSION
640 Capture enrichment is a promising alternative to PCR enrichment for metabarcoding. Its main
641 advantage is to provide better species detection thanks to its robustness to mismatches. At this point,
642 while performing much better than PCR enrichment, absolute biomass reconstruction is not applicable
643 without species mitogenome copy number correction and is therefore not tractable for community
644 studies where hundreds of species can be encountered. Yet, biomass rank ordination appears to be robust

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and could be used for ecological purpose. Albeit more bacterial contamination, the use of etDNA coupled to capture enrichment presents interesting compromises. It saves a lot of time by ignoring the organism picking step and permits to save organisms for further analyses. However, etDNA should be used for studies where quantitative information is not required. Finally, these promising results call for additional testing of the bait design strategy and hybridization reaction parameters, as well as testing capture enrichment on complex field samples.

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DATA ACCESSIBILITY

Additional reference sequences developed for bait design are available on Genbank (Accession Numbers MK584300:MK584515). COI and 16S sequences barcoded for the 10 species MC assignment

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960 are available on GenBank (Accession Numbers MK584516:MK584524 for COI and
961 MK584525:MK584534 for 16S). The bait set designed for this study is available on Zenodo (Zenodo
962 (https://doi.org/10.5281/zenodo.2581410). FASTQ files from PCR and capture enrichment are available
963 on NCBI (Accession Number SRA SRP188737). Scripts for bioinformatic analysis are available on
964 GitHub (https://github.com/mailysgauthier/bioinf-cap-PCR).
965
966 AUTHOR CONTRIBUTIONS
967 Samples of the 10 species MC were collected by MG. Methodology design was conceived by MG,
968 TL, CD, VE and TD. Laboratory work was conducted by MG, LKD, AN. Data analysis was conducted
969 by MG, TL and VE. MG and TL led the writing of the manuscript. All authors contributed to write the
970 manuscript.
971
972 TABLE AND FIGURE CAPTIONS
973 Table 1. Dry biomass in mg and number of individuals (italic) of freshwater invertebrate species in
974 the 10 species mock communities.
975
976 Table 2. Assessment of PCR and capture enrichment specificity on bulk and ethanol DNA. "% of
977 targeted reads": for capture enrichment only, percentage of reads that align to the COI, 16S, NAD1,
978 NAD4, NAD5, CYTB, and ATP6 reference databases, "% of COI reads": percentage of reads that align
979 to the COI protein reference database, "% of COI reads assigned to a MC species": percentage of the
980 number of reads that were successfully assigned to a species using the COI nucleotide reference
981 databases to the COI assigned reads and "% of non-protostomian reads": percentage of reads that align
982 to non-protostomian groups. Indicated values are mean number and standard deviation. For sample
983 values, see Table S4 (PCR enrichment) and Table S5 (Capture enrichment). 10 sps: 10 species MC; 52
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994 Figure 1: Overview of the experimental design of the study. Two different mock communities (MCs)
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Figure 4: Taxa recovery performance assessed using read proportion for etDNA of the 10 species mock communities. Differences between PCR (top left) and capture (top right) enrichment were assessed using a first ethanol extraction (extraction I). Differences between Fwh1 (bottom left) and BF2/BR2 (bottom right) primer pairs were assessed using a second ethanol extraction (extraction II). Sequence abundances are shown for each taxa (in rows) and mock community (in column). A crossed out cell indicates no assigned read. For detailed read counts, see Tables S6.

Figure 5: Principal Coordinate Analysis based on dissimilarity (Jaccard distances) in read proportions of each pair of the 10 species mock community samples. Samples are classified following their DNA template (etDNA and bulk DNA) and enrichment methods (BF2/BR2 amplicon, Fwh1 amplicon and capture). etDNA1 and etDNA2 correspond to the first and second extraction of DNA from ethanol, respectively. Percentage of variance explained by each axis is shown in bracket. Ellipses are drawn with a confidence limit of 0.95. Samples are grouped by enrichment method rather than by DNA template.

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Table 1. Dry biomass in mg and number of individuals (italic) of freshwater invertebrate species in the 10 species mock communities.

	Mock community							
Species	1	2	3	4	5	6	7	8
Chironomus riparius Meigen, 1804	1.84 *(5)*	1.83 *(5)*	2.58 *(5)*	2.42 *(5)*	3.66 *(10)*	4.62 *(11)*	3.89 *(10)*	4.44 *(10)*
Epeorus assimilis Eaton, 1885	102.32 *(5)*	55.53 *(6)*	77.89 *(5)*	96.47 *(5)*	62.94 *(4)*	110.73 *(4)*	114.80 *(4)*	74.23 *(4)*
Heptagenia sulphurea (O.F. Müller, 1776)	15.73 *(5)*	20.79 *(5)*	21.08 *(5)*	18.02 *(5)*	38.31 *(10)*	28.19 *(10)*	32.58 *(10)*	38.37 *(10)*
Isoperla rivulorum (Pictet, 1841)	37.97 *(5)*	31.88 *(5)*	36.28 *(5)*	30.89 *(5)*	35.11 *(6)*	33.39 *(6)*	36.22 *(6)*	39.47 *(6)*
Nemurella picteti Klapálek, 1900	2.83 *(5)*	2.49 *(5)*	2.34 *(5)*	3.60 *(5)*	0.50 *(1)*	0.97 *(1)*	0.77 *(1)*	0.56 *(1)*
Hydropsyche siltalai Doehler, 1963	68.42 *(5)*	53.15 *(5)*	53.63 *(5)*	53.32 *(5)*	80.36 *(8)*	66.27 *(8)*	86.71 *(8)*	79.86 *(8)*
Athripsodes aterrimus (Stephens, 1836)	6.10 *(6)*	6.68 *(5)*	7.87 *(5)*	6.23 *(5)*	9.10 *(6)*	7.96 *(6)*	5.39 *(6)*	7.01 *(6)*
Physella acuta (Draparnaud, 1805)	23.30 *(5)*	24.85 *(5)*	23.03 *(5)*	31.30 *(5)*	16.81 *(4)*	15.22 *(4)*	13.25 *(4)*	20.68 *(4)*
Ancylus fluviatilis O.F. Müller, 1774	13.42 *(5)*	18.16 *(5)*	18.40 *(5)*	15.42 *(5)*	20.79 *(8)*	20.47 *(8)*	23.21 *(8)*	15.98 *(8)*
Gammarus fossarum Koch, 1836	23.07 *(5)*	20.28 *(5)*	16.89 *(5)*	16.72 *(5)*	4.4 *(1)*	4.02 *(1)*	5.17 *(1)*	3.76 *(1)*

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bulk DNA	*Mock community*	*Raw reads*	*Quality filtered reads*	*% targeted reads*	*% COI reads*	*% of COI reads assigned to a MC species*	*% of non-protostomian reads (contaminants)*
PCR enrichment	10 sps	170,026 ± 40,272	114,701 ± 31,160	Ø	49.54 ± 4.83	96.01 ± 3.24	2.77 ± 1.30
	52 taxa	135,168 ± 40,871	91,935 ± 32,720	Ø	42.33 ± 4.61	96.55 ± 3.80	2.37 ± 1.36
Capture enrichment	10 sps	493,967 ± 288,240	243,483 ± 86,526	43.34 ± 24.52	25.16 ± 18.81	61.17 ± 5.55	10.73 ± 3.78
	52 taxa	553,825 ± 165,868	264,694 ± 75,864	80.21 ± 5.09	69.30 ± 6.51	63.27 ± 1.96	5.31 ± 1.70
Enrichment-free	10 sps	921,863 ± 68,162	574,443 ± 43,547	1.19 ± 0.31	0.023 ± 0.028	26.14 ± 9.21	31.66 ± 1.96
etDNA	*Mock community*	*Raw reads*	*Quality filtered reads*	*% targeted reads*	*% COI reads*	*% of COI reads assigned to a MC species*	*% of non-protostomian reads (contaminants)*
PCR enrichment - Fwh1 extraction I	10 sps	283,644 ± 52,944	233,856 ± 43,256	Ø	62.97 ± 4.47	46.36 ± 11.47	18.61 ± 12.78
PCR enrichment - Fwh1 extraction II	10 sps	205,229 ± 81,671	169,772 ± 69,282	Ø	66.34 ± 5.88	47.72 ± 14.65	10.13 ± 4.12
PCR enrichment - BF2/BR2 extraction II	10 sps	203,406 ± 43,493	123,148 ± 47,936	Ø.	48.97 ± 9.81	60.95 ± 18.88	28.21 ± 19.29
Capture Enrichment – extraction I	10 sps	517,202 ± 234,942	309,637 ± 147,304	27.18 ± 8.39	19.09 ± 5.82	33.71 ± 10.96	56.81 ± 10.17

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			Random effect		
bulk DNA	*Fixed effect*	*Random effect*	*None*	*Intercept*	*Intercept + slope*
PCR enrichment	1	Species	2,012,800	31,216	Ø
		Sample		2,012,802	
	biomass proportion	Species	537,148	6,674	0
		Sample		496,237	382,656
Capture enrichment	1	Species	320,984	9,223	Ø
		Sample		320,986	
	biomass proportion	Species	185,777	5,154	0
		Sample		185,322	182,666
etDNA	*Fixed effect*	*Random effect*	*None*	*Intercept*	*Intercept + slope*
PCR enrichment	1	Species	2,602,013	121,099	Ø
		Sample		2,602,015	
	biomass proportion	Species	1,897,689	97,590	0
		Sample		1,894,974	1,748,461
Capture enrichment	1	Species	195,670	43,136	Ø
		Sample		195,672	
	biomass proportion	Species	166,761	26,717	0
		Sample		166,760	102,699

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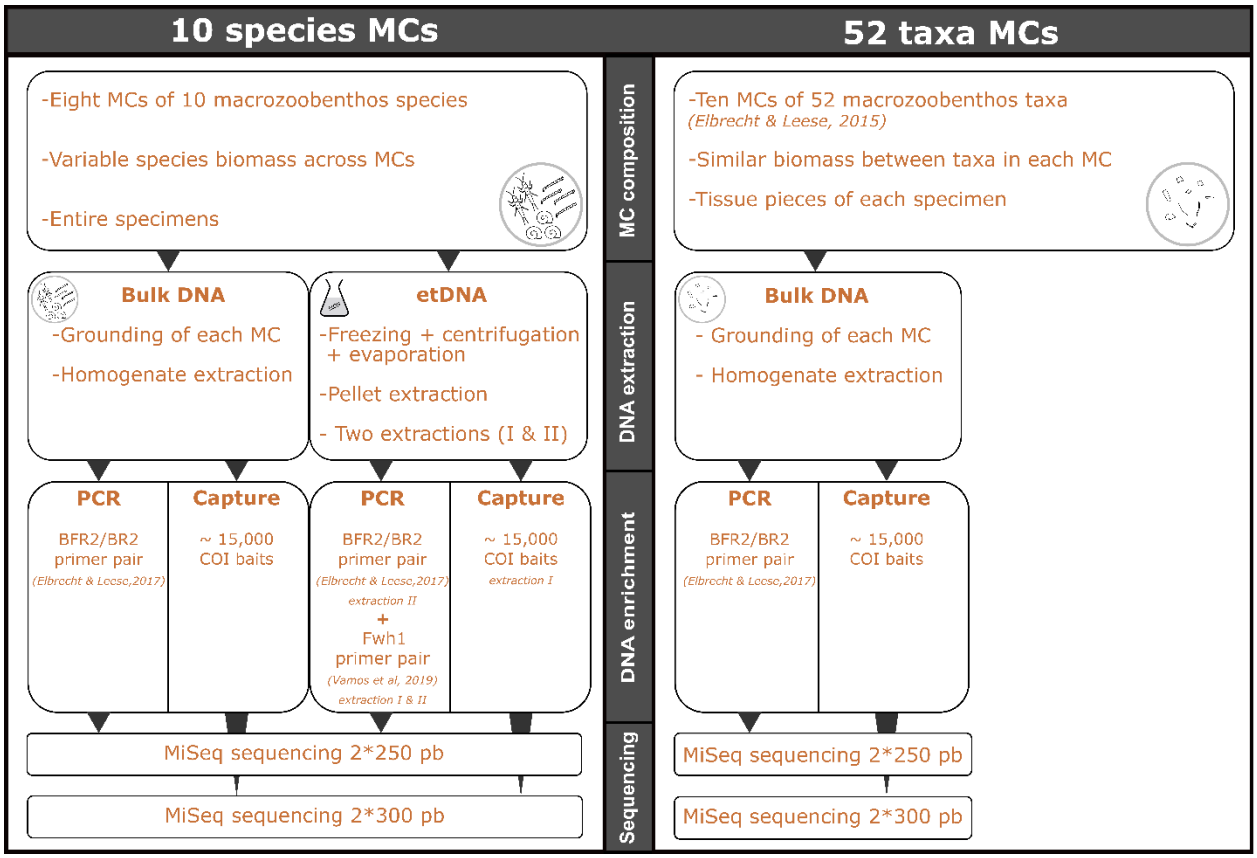

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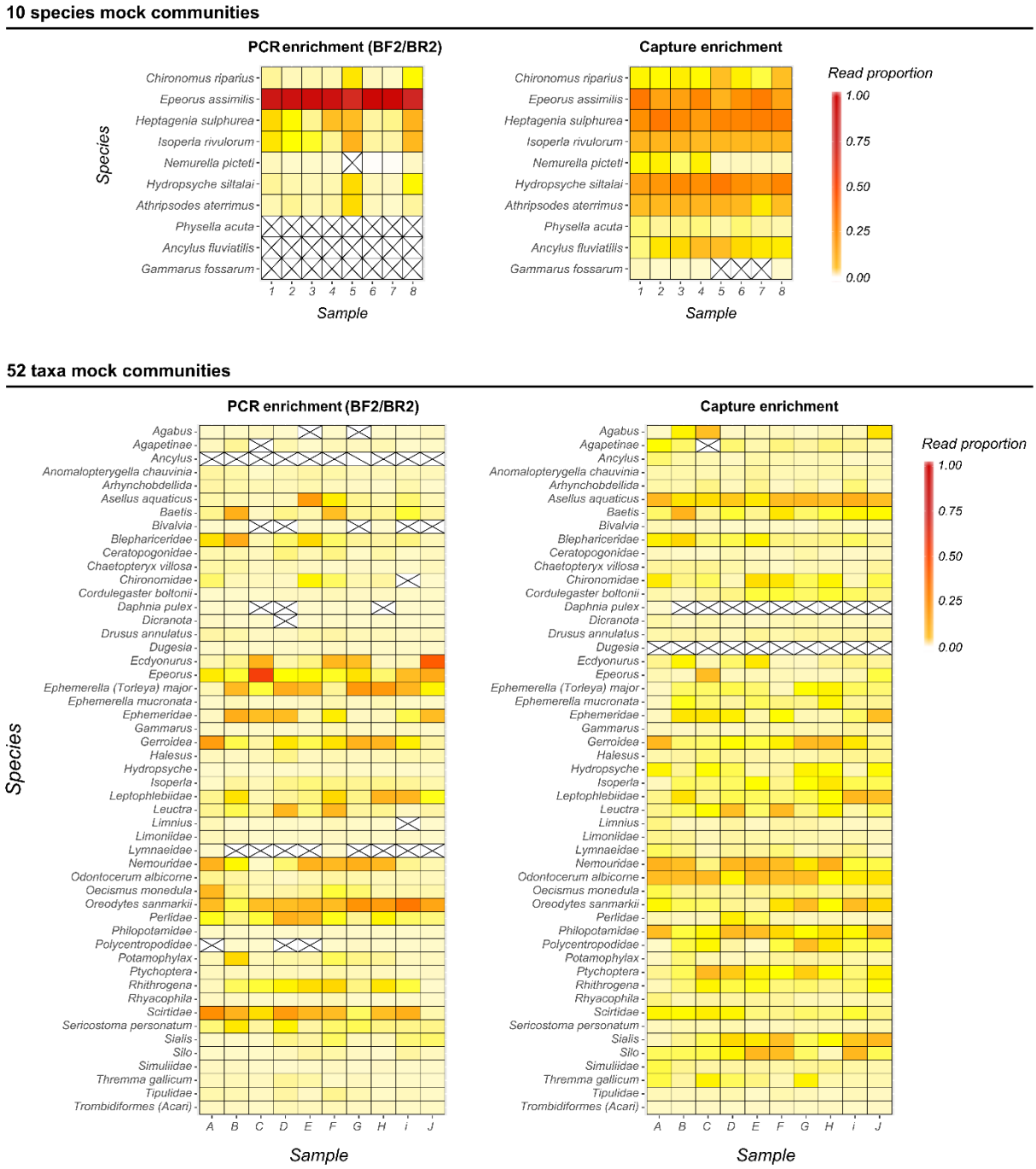

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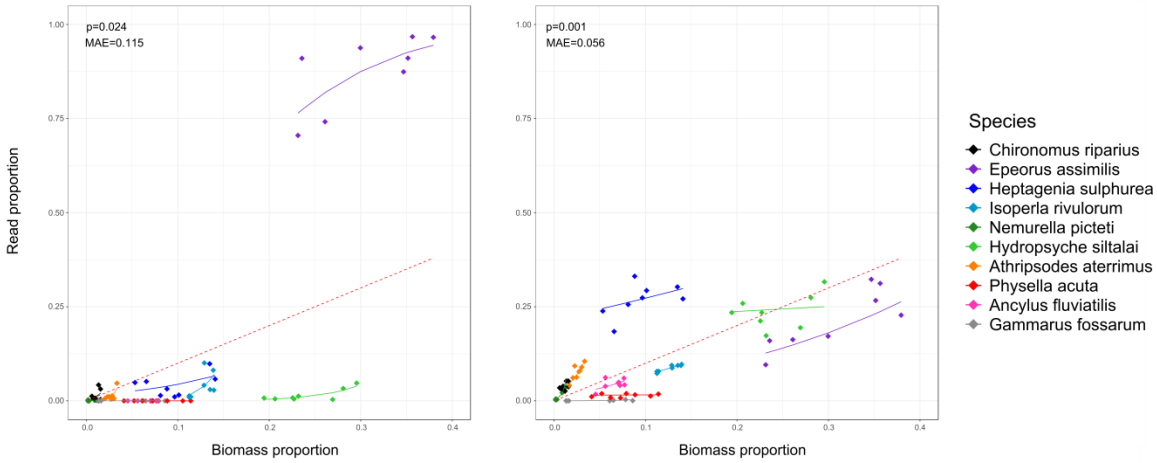

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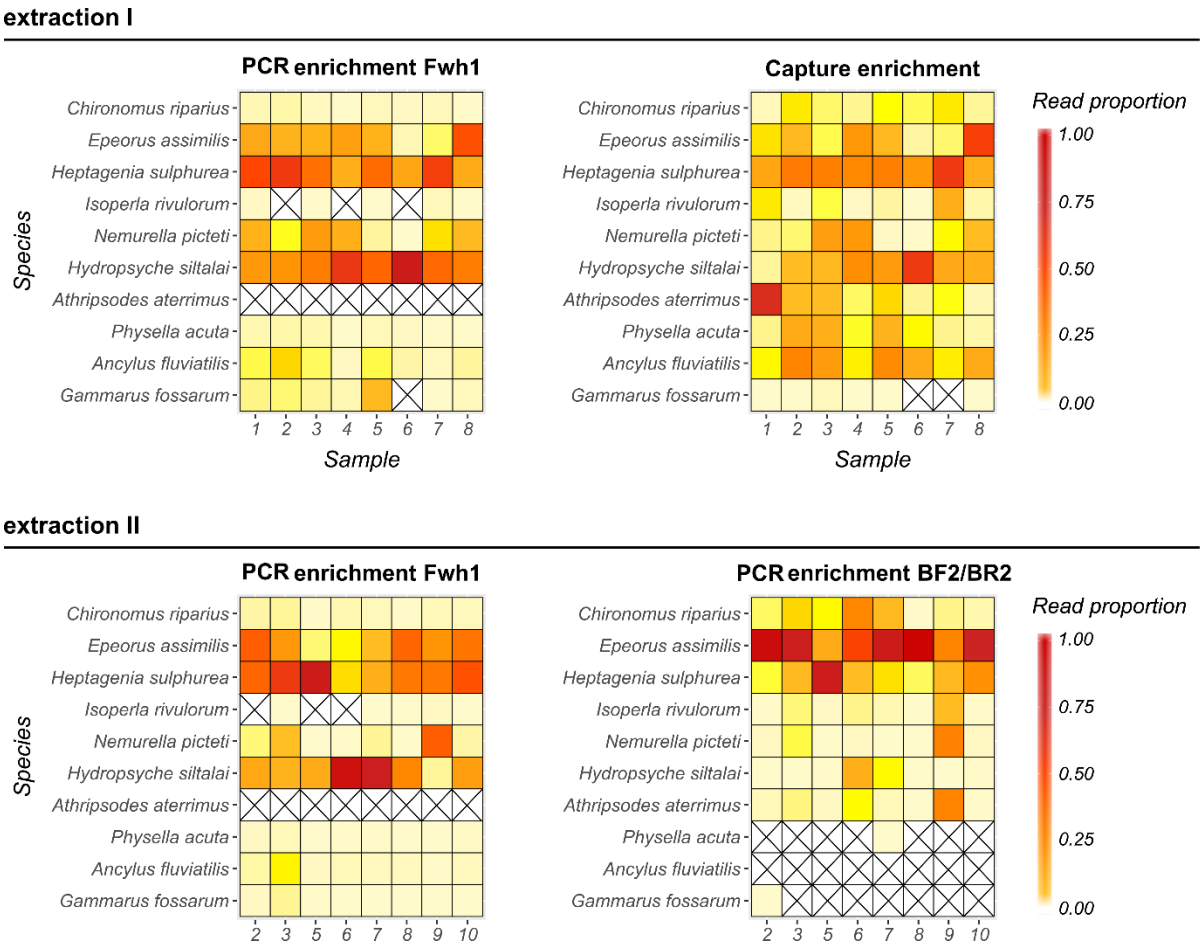

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