



HAL
open science

Experimental assessment of a new comprehensive model for single ring infiltration data

Simone Di Prima, Mirko Castellini, Majdi Abou Najm, Ryan Stewart, Rafaël Angulo-Jaramillo, Thierry Winiarski, Laurent Lassabatère

► **To cite this version:**

Simone Di Prima, Mirko Castellini, Majdi Abou Najm, Ryan Stewart, Rafaël Angulo-Jaramillo, et al.. Experimental assessment of a new comprehensive model for single ring infiltration data. *Journal of Hydrology*, 2019, 573, pp.937-951. 10.1016/j.jhydrol.2019.03.077 . hal-02141637

HAL Id: hal-02141637

<https://univ-lyon1.hal.science/hal-02141637>

Submitted on 22 Oct 2021

HAL is a multi-disciplinary open access archive for the deposit and dissemination of scientific research documents, whether they are published or not. The documents may come from teaching and research institutions in France or abroad, or from public or private research centers.

L'archive ouverte pluridisciplinaire **HAL**, est destinée au dépôt et à la diffusion de documents scientifiques de niveau recherche, publiés ou non, émanant des établissements d'enseignement et de recherche français ou étrangers, des laboratoires publics ou privés.



Distributed under a Creative Commons Attribution - NonCommercial 4.0 International License

1 Experimental Assessment of a New Comprehensive Model for Single Ring Infiltration Data

2 Simone Di Prima ^{a,*}, Mirko Castellini ^b, Majdi R. Abou Najm ^c, Ryan D. Stewart ^d, Rafael Angulo-Jaramillo

3 ^a, Thierry Winiarski ^a and Laurent Lassabatere ^a

4 ^a Université de Lyon; UMR5023 Ecologie des Hydrosystèmes Naturels et Anthropisés, CNRS, ENTPE, Université Lyon 1, 3 rue
5 Maurice Audin, 69518 Vaulx-en-Velin, France.

6 ^b Council for Agricultural Research and Economics-Agriculture and Environment Research Center (CREA-AA), Via Celso Ulpiani
7 5, 70125 Bari, Italy.

8 ^c Department of Land, Air and Water Resources, University of California, Davis, CA, 95616, United States.

9 ^d School of Plant and Environmental Sciences, Virginia Polytechnic Institute and State University, Blacksburg, VA, United States.

10 * Corresponding Author. E-mail: simone.diprima@entpe.fr

11 Abstract

12 The objective of this paper was to evaluate a recently proposed comprehensive model for three-
13 dimensional single-ring infiltration and its suitability for estimating soil hydraulic properties.
14 Infiltration data from four different soils with contrasting characteristics were inverted to estimate
15 field-saturated soil hydraulic conductivity, K_{fs} , values using a total of fourteen different scenarios.
16 Those scenarios differed by: i) the way they constrained the macroscopic capillary length, λ , and the
17 initial and saturated soil water contents, θ_i and θ_s , ii) the use of transient or steady-state data, and iii)
18 the fitting methods applied to transient data. For comparative purposes, the SSBI method (Steady
19 version of the Simplified method based on a Beerkan Infiltration run) was also applied. For
20 validation purposes K_{fs} data estimated from the different scenarios were compared with those values
21 obtained by numerical inverse modeling with HYDRUS-2D/3D. This comparison identified
22 Approaches 1 and 3, which respectively estimate K_{fs} via optimization and using analytical
23 expressions, as the most accurate methods. The steady-state scenario of Approach 4 and the SSBI
24 method, both of which use a λ value of first approximation, appeared preferable for field campaigns
25 aimed to sample remote or large areas, given that they do not need additional data and still provide
26 acceptable estimates. The reliability of K_{fs} data was also checked through a comparison with
27 unsaturated hydraulic conductivity, K_h , values measured in laboratory on extracted soil cores, in
28 order to discriminate between theoretically possible ($K_{fs} > K_h$) and impossible ($K_{fs} \leq K_h$) situations.
29 Physically possible K_{fs} values were always obtained with the exception of the crusted soil, where K_{fs}
30 $< K_h$ situations suggested that the crust layer reduced water flow during ponding experiments in the

31 field. The new comprehensive model tested in this study represents a valuable tool for analyzing
32 both transient and steady-state infiltration data, as well as experiments carried out with different
33 depths of ponded water, ring sizes and ring insertion depths.

34 **Keywords:** infiltration model, single-ring infiltrometer, Beerkan, hydraulic conductivity.

35 **1. Introduction**

36 Knowledge of soil properties is essential for modeling hydrological processes. Among other
37 properties, the field-saturated soil hydraulic conductivity, K_{fs} , has an important role in the
38 partitioning of rainfall into runoff and infiltration (Dusek et al., 2012). Different devices and
39 techniques have been developed over time to measure K_{fs} in the field, such as the Guelph
40 permeameter, the double- and the single-ring infiltrometers, among others (Angulo-Jaramillo et al.,
41 2016). The Guelph permeameter is a device that establishes three-dimensional, constant-head
42 infiltration within a small well excavated into the soil (Reynolds and Elrick, 1985). The double-ring
43 infiltrometer uses two concentric rings, namely an inner ring and a buffering ring, to create a one-
44 dimensional (1D) infiltration process under the inner ring (Reynolds et al., 2002). However, some
45 limitations may be encountered in the field when applying these methods. When using the Guelph
46 permeameter, the excavation of the well may cause soil compaction, artificially decreasing the
47 infiltration rates (Bagarello et al., 1999). The water flow under the inner ring of the double-ring
48 infiltrometer rarely approaches a one-dimensional infiltration process in practice (Reynolds et al.,
49 2002). Moreover, this latter method also requires a large amount of water to maintain ponding
50 conditions inside the buffering ring, thus limiting its application in remote areas.

51 The single-ring infiltrometer technique (Reynolds and Elrick, 1990) is a widespread method
52 (e.g., Braud et al., 2017), which has the advantage of speed and simplicity over more cumbersome
53 procedures, such as the Guelph permeameter and the double-ring infiltrometer. With a single-ring
54 infiltrometer, a constant or falling-head infiltration process has to be established. Different methods

55 for calculating K_{fs} from single-ring data have been developed over time. Among them, the one-
56 ponding depth method by Reynolds and Elrick (1990) and the similar method by Wu et al. (1999)
57 both estimate K_{fs} from steady-state single-ring infiltrometer data. Other approaches make use of
58 transient infiltration data (e.g., Wu et al., 1999; Wu and Pan, 1997) to determine K_{fs} . These
59 alternative approaches may alleviate the experimental efforts needed to determine K_{fs} data in the
60 field (Di Prima et al., 2018b). For instance, limiting the analysis to the transient phase may prove
61 advantageous when characterizing low permeability soils, by reducing the required measurement
62 time (Bagarello et al., 2014c).

63 A variation of the single-ring infiltrometer technique is the Beerkan experiment, which consists
64 of infiltrating water through a ring inserted shallowly (e.g., 1 cm) into the soil with a quasi-zero
65 head of water imposed on the soil surface (Braud et al., 2005). Many different methods have been
66 used to interpret Beerkan data. As an example, the Beerkan Estimation of Soil Transfer parameters
67 (BEST) methods (Bagarello et al., 2014b; Lassabatere et al., 2006; Yilmaz et al., 2010) enable the
68 user to derive the whole set of soil hydraulic parameters related to water retention and unsaturated
69 hydraulic conductivity curves. Bagarello et al (2014c) and Bagarello et al. (2017) proposed the
70 TSBI and SSBI methods (i.e., the Transient and Steady Simplified methods based on a Beerkan
71 Infiltration run), which allow to estimate K_{fs} by only using a Beerkan experiment.

72 In terms of drawbacks, BEST methods require collection of supplementary data, e.g., bulk
73 density, particle size distribution. The models also typically fit only transient or steady-state data,
74 but not both. Such peculiarities make it difficult or impossible to apply these methods to
75 heterogeneous datasets, such as the recently developed Soil Water Infiltration Global (SWIG)
76 database (Rahmati et al., 2018). Recently, Stewart and Abou Najm (2018a) developed a new
77 comprehensive model for single ring infiltration data by combining the infiltration models by
78 Reynolds and Elrick (1990) and Wu et al. (1999). These authors proposed four different approaches
79 for estimating K_{fs} values from both transient and steady-state single-ring infiltration data. The four
80 approaches differ in the way they constrain the macroscopic capillary length, λ , and the initial and

81 saturated soil water contents, θ_i and θ_s ; each approach requires different types of input parameters
 82 and exhibits different types and amounts of error. The proposed model has a practical interest in
 83 that it treats both transient and steady-state infiltration data, and can analyze experiments carried out
 84 with different ring sizes and ring insertion depths. However, the model was previously validated
 85 using only laboratory and numerical experiments, meaning that it has not yet been experimentally
 86 validated with field measurements.

87 The objective of this research was to test this new comprehensive model (Stewart and Abou Najm,
 88 2018a) using data acquired for four soils with a range of physical and hydraulic properties. The
 89 model estimated K_{fs} using the four different approaches for constraining λ , θ_i and θ_s , along with
 90 several methods for determining infiltration constants, for a total of thirteen scenarios. The SSBI
 91 method developed by Bagarello et al. (2017) was also applied, giving a fourteenth scenario. The
 92 reliability of K_{fs} estimates were verified first through a comparison with laboratory measurements of
 93 unsaturated hydraulic conductivity, and then via comparison with values obtained by numerical
 94 inverse modeling with HYDRUS-2D/3D.

95 2. Theory

96 2.1. Analysis of single-ring infiltrometer data

97 The model proposed by Stewart and Abou Najm (2018a) describes three dimensional (3D)
 98 cumulative infiltration, I (L), from a surface circular source under a positive pressure head using the
 99 following explicit relationships for transient and steady-state conditions:

$$100 \quad I = \sqrt{\frac{(\theta_s - \theta_i) (h_{source} + \lambda) K_{fs}}{b}} \sqrt{t} + afK_{fs}t \quad t < \tau_{crit}$$

101 (1a)

$$102 \quad I = \frac{(\theta_s - \theta_i) (h_{source} + \lambda) K_{fs}}{4fb(1-a)} + fK_{fs}t \quad t \geq \tau_{crit}$$

103 (1b)

104 where t (T) is the time, τ_{crit} (T) is the maximum time for which the transient relationship can be
 105 considered valid, θ_s (L^3L^{-3}) and θ_i (L^3L^{-3}) are respectively the saturated and initial volumetric soil
 106 water content, h_{source} (L) is the established ponding depth of water, λ (L) is the macroscopic
 107 capillary length of the soil, K_{fs} ($L T^{-1}$) is the field-saturated soil hydraulic conductivity, a and b are
 108 dimensionless constants respectively equal to 0.45 and 0.55, and f is a correction factor that depends
 109 on soil initial and boundary conditions and ring geometry (Reynolds and Elrick, 1990):

$$110 \quad f = \frac{h_{source} + \lambda}{G^*} + 1 \quad (2)$$

111 in which the G^* (L) term is equal to:

$$112 \quad G^* = d + \frac{r_d}{2} \quad (3)$$

113 where d (L) is the ring insertion depth into the soil and r_d (L) is the radius of the disk source.

114 Because τ_{crit} is not known a priori, the criterion suggested by Bagarello et al. (1999) can be
 115 considered to discriminate between transient and steady-state conditions for cumulative infiltration
 116 data. Assuming the steady-state conditions are reached before the end of an infiltration run, a linear
 117 regression analysis is conducted for the last three data points of $I(t)$ versus t . The time to steady-
 118 state, t_s (L), is determined as the first value for which:

$$119 \quad \hat{E} = \left| \frac{I(t) - I_{reg}(t)}{I(t)} \right| \times 100 \leq E \quad (4)$$

120 where $I_{reg}(t)$ is estimated from regression analysis, and E defines a given threshold to check
 121 linearity. Equation (4) is applied from the start of the experiment until finding the first data point
 122 that fits the condition $\hat{E} \leq E$ (Angulo-Jaramillo et al., 2016). An illustrative example of t_s
 123 estimation using the commonly used value of $E = 2$ is shown in **Figure 1a**. Transient infiltration
 124 conditions therefore occur from time 0 until time t_s (i.e., when $\hat{E} > 2$; **Figure 1a**), while steady-
 125 state conditions exist for all data points measured after time t_s (i.e., when $\hat{E} \leq 2$).

126 Equations (1) can be simplified as follows (Philip, 1957):

127
$$I = c_1\sqrt{t} + c_2t \quad (5a)$$

128
$$I = c_3 + c_4t \quad (5b)$$

129 where the intercept, c_3 (L), and the slope, c_4 (L T⁻¹), are estimated by linear regression analysis of
 130 the $I(t)$ vs. t plot, while the infiltration coefficients c_1 (L T^{-0.5}) and c_2 (L T⁻¹) can be determined
 131 according to the fitting methods referred to as cumulative infiltration (CI, e.g. Zhang, 1997),
 132 cumulative linearization (CL, Smiles and Knight, 1976) and differential linearization (DL,
 133 Vandervaere et al., 1997). In this investigation we considered all three fitting methods, since each
 134 method has its own advantages and peculiarities (Vandervaere et al., 2000a). An example of the
 135 fitting procedures is depicted in **Figures 1b, c, d**.

136 **2.2. Estimation of field-saturated soil hydraulic conductivity values**

137 Stewart and Abou Najm (2018b) proposed four different approaches, named Approaches 1, 2, 3
 138 and 4, for estimating K_{fs} values from single-ring infiltration data. The differences between the four
 139 approaches involve the way in which λ , θ_i and θ_s are constrained, which must occur before
 140 estimating K_{fs} . In the following sections, the four approaches are briefly explained.

141 **2.2.1. Approach 1**

142 The first approach estimates K_{fs} by constraining all of the other considered parameters, i.e., λ , θ_i
 143 and θ_s , and then fitting Eq. (1) to cumulative infiltration. Stewart and Abou Najm (2018b) proposed
 144 to estimate λ from water retention data. Specifically, according to these authors, if the soil is
 145 relatively dry at the beginning of the infiltration experiment, λ tends towards a maximum value, λ_{max}
 146 (L), defined as:

147
$$\lambda_{max} = \frac{h_b\eta}{1-\eta} \quad (6)$$

148 where η and h_b (L) are respectively the pore size index and the head scale parameter of the Brooks
 149 and Corey (1964) relations for water retention and hydraulic conductivity. Note that Eq. (6) can be

150 considered valid for values of the initial matric head of the soil, h_i (L), ranging between $-\infty$ and $2h_b$
 151 (Stewart and Abou Najm, 2018a). Initial and saturated volumetric soil water contents (θ_i and θ_s)
 152 may be measured from soil samples collected before and after the infiltration run or otherwise
 153 estimated.

154 2.2.2. Approach 2

155 Approach 2 only requires estimates for θ_i and θ_s . For transient-state data, once the c_1 and c_2
 156 coefficients are determined, the field-saturated soil hydraulic conductivity and the macroscopic
 157 capillary length are calculated by the following equations:

$$158 \quad K_{fs} = \frac{c_2}{a} - \frac{bc_1^2}{(\theta_s - \theta_i) G^*} \quad (7)$$

$$159 \quad \lambda = \frac{bc_1^2}{K_{fs}(\theta_s - \theta_i)} - h_{source} \quad (8)$$

160 While for steady-state data, K_{fs} and λ are calculated as:

$$161 \quad K_{fs} = \frac{c_4 G^*}{\lambda + h_{source} + G^*} \quad (9)$$

$$162 \quad \lambda = \frac{4c_3 b(1-a)(h_{source} + G^*) - h_{source}(\theta_s - \theta_i) G^*}{(\theta_s - \theta_i) G^* - 4c_3 b(1-a)} \quad (10)$$

163 2.2.3. Approach 3

164 This approach allows the estimation of the field-saturated soil hydraulic conductivity using only
 165 λ , estimated by Eq. (6), and c_2 or c_4 , as determined from the infiltration run. For transient-state data,
 166 once λ and c_2 are established then the field-saturated soil hydraulic conductivity is calculated by the
 167 following equation:

$$168 \quad K_{fs} = \frac{c_2}{a \left(\frac{h_{source} + \lambda}{G^*} + 1 \right)} \quad (11)$$

169 While for steady-state data, K_{fs} is calculated as:

170
$$K_{fs} = \frac{c_4}{\left(\frac{h_{source} + \lambda}{G^*} + 1 \right)} \quad (12)$$

171 **2.2.4. Approach 4**

172 Approach 4 uses Eqs. (11) and (12) in conjunction with a λ value of first approximation.
 173 Following Stewart and Abou Najm (2018b), a value of $\lambda = 150$ mm was selected for this
 174 investigation. This approach does not require additional information to estimate K_{fs} from infiltration
 175 runs. Therefore, it is particularly useful when a large number of locations needs to be sampled,
 176 particularly when time and financial resources are limited.

177 **2.2.5. SSBI method**

178 For comparative purposes, the SSBI method (Steady-state version of the Simplified method
 179 based on a Beerkan Infiltration run) proposed by Bagarello et al. (2017) was also applied to
 180 estimate K_{fs} . SSBI estimates K_{fs} through a Beerkan infiltration test, i.e., a simple 3D infiltration run
 181 with a quasi-zero water pressure head at the soil surface (Braud et al., 2005; Lassabatere et al.,
 182 2006), by the following equation:

183
$$K_{fs} = \frac{c_4}{\frac{1.364 \lambda}{r_d} + 1} \quad (13)$$

184 Note that Eq. (13) is analogous to Eq. (16) in Bagarello et al. (2017), with the latter considering
 185 the sorptive number, α^* (L^{-1}), which is equal to λ^{-1} (Angulo-Jaramillo et al., 2016). Because Eqs.
 186 (12) and (13) are analogous to one another (i.e., both require estimates for c_4 and λ to determine
 187 K_{fs}), the SSBI method was also applied assuming $\lambda = 150$ mm.

188 3. Material and methods

189 3.1. Soil Sampling

190 Four soils with contrasting physical and hydraulic properties were evaluated in this study
191 (Castellini et al., 2018). According to the USDA classification, a sandy soil was sampled at Arborea
192 in Sardinia and a silty-loam soil was sampled at the experimental farm of CREA-AA in Foggia,
193 Apulia. Two sandy-loam soils were sampled in Sicily at the Department of Agriculture, Food and
194 Forest Sciences of the Palermo University (sandy-loam 1) and Villabate (sandy-loam 2). For each
195 site, a total 10 undisturbed soil cores (50 mm in height and 50 mm in diameter) were collected at
196 randomly sampled points and used to determine both the soil bulk density, ρ_b (g cm^{-3}), and the
197 initial volumetric soil water content, θ_i ($\text{cm}^3\text{cm}^{-3}$). The soil porosity was calculated from the ρ_b data,
198 assuming a soil particle density of 2.65 g cm^{-3} . The field saturated soil water content, θ_s ($\text{cm}^3 \text{ cm}^{-3}$),
199 was considered equal to the porosity, in line with other studies (e.g., Di Prima et al., 2018d;
200 Mubarak et al., 2009).

201 Disturbed soil samples were also collected to determine the particle size distribution. The
202 samples were air-dried and sieved through a 2-mm mesh. H_2O_2 pretreatment was used to eliminate
203 organic matter and clay deflocculation was encouraged using sodium metaphosphate and
204 mechanical agitation (Gee and Bauder, 1986). Fine size fractions were determined by the
205 hydrometer method, whereas the coarse fractions were obtained by mechanical dry sieving. The soil
206 organic carbon content, SOC (%), was determined by the Walkley–Black method (Walkley and
207 Black, 1934). Then, the soil organic matter content, SOM (%), was estimated using the van
208 Bemmelen conversion factor of 1.724 (Van Bemmelen, 1890). The measured soil physical
209 properties are summarized in **Table 1**. Furthermore, five to nine undisturbed soil cores (85 mm in
210 diameter by 75 mm in height) were also collected at each sampling site to conduct measurements of
211 unsaturated hydraulic conductivity and evaporation tests in the laboratory.

212 **3.2. Laboratory measurements of unsaturated hydraulic conductivity**

213 Laboratory measurements of unsaturated hydraulic conductivity, K_h , were collected to verify the
214 reliability of K_{fs} estimates. We chose to use unsaturated as opposed to saturated conditions as a way
215 to minimize uncertainty due to measurement artifacts such as entrapped air, open-ended pores, and
216 edge flow. All these phenomena can result in considerable variations between field-measured and
217 laboratory-derived estimates of hydraulic conductivity (Di Prima et al., 2018c; Sakaguchi et al.,
218 2005; Stewart and Abou Najm, 2018b). Obviously, the above-mentioned uncertainties are expected
219 to be less noticeable or even negligible for unsaturated measurements. In particular, the comparison
220 between K_{fs} and K_h data allowed us to discriminate between possible ($K_{fs} > K_h$) and physically
221 impossible ($K_{fs} \leq K_h$) situations. The unit hydraulic gradient method (Klute and Dirksen, 1986) was
222 used to determine the unsaturated soil hydraulic conductivity, K_h (mm h^{-1}), on the 85 mm by 75 mm
223 soil cores. According to the procedure described by Bagarello et al. (2007) and Castellini et al.
224 (2015), the upper layer of the soil (≤ 2 mm) was carefully removed to allow the placement of a
225 nylon guard cloth with an air entry value of -160 mm and a thin contact material layer (Spheriglass,
226 glass spheres, no. 2227). The nylon guard cloth was also placed at the bottom face of the sample to
227 avoid soil displacement. Each sample was positioned on a sintered porous plate having an air entry
228 value of -400 mm and then connected to an outflow tube that could be moved in height to establish
229 a given pressure head value at the bottom of the core. The sample was previously equilibrated for a
230 48 h time interval on the porous plate by repeatedly raising the outflow level at the first pressure
231 head value (-120 or -75 mm, depending on the sample). A negative pressure head at the top of the
232 sample, h_0 , was imposed by the tension infiltrometer device, which consisted of a porous disk (85
233 mm in diameter) connected to the water supply reservoir. Measurements were performed by
234 applying the same pressure head value at the two ends of the soil core. Infiltration evolved from an
235 initial transient stage to a steady-state stage in which a unit hydraulic gradient was obtained (i.e.,
236 infiltration rates were constant and pressure head readings were equal throughout the soil core). For

237 this stage, the steady-state flux was equivalent to the unsaturated hydraulic conductivity
238 corresponding to the imposed pressure head value (Bagarello et al., 2007). For the sandy, sandy-
239 loam 1 and sandy-loam 2 samples, the pressure head sequences applied was $h_0 = -120, -60, -30,$ and
240 -10 mm, whereas for the silty-loam samples, the sequence was $h_0 = -75, -30,$ and -10 mm.

241 **3.3. Laboratory evaporation experiments**

242 Soil water retention measurements were carried out on the same undisturbed soil cores used to
243 run the unit hydraulic gradient experiments. These experiments allowed us to optimize the
244 parameters of the Brooks and Corey (1964) relationship. In this way, an independent estimate of the
245 macroscopic capillary length (required for Approaches 1 and 3) was determined by inputting the
246 shape, η , and scale, h_b (L), parameters into Eq. (6). In this investigation, we used the evaporation
247 method proposed by Wind (1969) for the computation of the water retention curve, $\theta(h)$, through
248 the simultaneous measurement of volumetric soil water contents and pressure heads at multiple
249 depths during an evaporation process. More details on the laboratory procedure can be found in
250 Castellini et al (2018). The fitting of the water retention data was performed using the program
251 SWRC Fit developed by Seki (2007). This program uses an iterative nonlinear regression procedure
252 that finds the values of the optimized parameters by minimizing the sum of the squared residuals
253 between the model and the observed data. Parameter values are reported in **Table 2**.

254 **3.4. Ponding infiltrometer runs**

255 For each site, a total of ten ponded infiltration runs of the Beerkan type (Braud et al., 2005;
256 Lassabatere et al., 2006) were carried out at different sampling points. According to the existing
257 literature, the chosen sample size ($N = 10$) was expected to yield representative mean K_{fs} values at
258 the field scale (Reynolds et al., 2000; Verbist et al., 2010). A ring with an inner diameter of 150 mm
259 was used in the Apulian (Foggia, silty-loam) and Sardinian (Arborea, sandy) sites, and a ring with
260 an inner diameter of 85 mm was used in both of the Sicilian sites (sandy-loam 1 and 2). At the

261 Arborea site, a larger ring diameter was chosen due to the presence of a weak but clearly detectable
262 surface structural crust (thickness of ~2 mm), due to the possibility that fractures along the ring
263 edge may connect the ponded surface water with the underlying, non-cruste, soil layer
264 (Vandervaere et al., 1997). Beyond the large ring diameter, the small insertion depth (10 mm) used
265 in this study should also help mitigate the formation of fractures through the crusted layer during
266 ring insertion (Alagna et al., 2019; Souza et al., 2014).

267 As prescribed by the Beerkan experimental procedure, the ring was inserted to a depth of 10 mm
268 in all sites. For each run, 15 water volumes, each equal to 64 mL for the 85 mm diameters rings and
269 200 mL for the 150 mm diameter rings, were successively poured on the confined soil surface. The
270 number of infiltrated volumes was sufficient to reach steady-state, as required by the Beerkan
271 method (Lassabatere et al., 2006). The energy of the falling water was dissipated with fingers to
272 minimize the soil disturbance owing to water pouring, as commonly suggested (e.g., Alagna et al.,
273 2016; Bagarello et al., 2014a). For each water volume, the time needed for the water to infiltrate
274 was recorded, and the cumulative infiltration, I (mm) was plotted against time, t (h).

275 **3.5. Numerical Simulation**

276 We chose to use the $K_{fs-HYDRUS}$ values obtained by the inverse procedure in HYDRUS-2D/3D
277 (Šimůnek et al., 2008) as a benchmark, as an independent K_{fs} datum that can be used for assessing
278 simplified procedures or validating new developed methods does not currently exist (Bagarello et
279 al., 2017). As discussed above, laboratory measurements induce experimental artifacts that may
280 limit their comparability with in-situ measurements (Di Prima et al., 2018c). Discrepancies are also
281 expected when different measurement techniques are applied in the field or even when the same
282 dataset is analyzed by alternative calculation approaches (Mertens et al., 2002), though in the latter
283 case the results can still be compared to one another (Wu et al., 1999). The inverse procedure using
284 in HYDRUS-2D/3D combines the Levenberg-Marquardt non-linear parameter optimization method
285 (Marquardt, 1963) with a numerical solution of the axisymmetric form of Richards equation

286 (Angulo-Jaramillo et al., 2000; Šimůnek and Hopmans, 2002). Similarly to the model configuration
287 used by Stewart and Abou Najm (2018b), the soil columns were modeled as a 2D axisymmetric
288 plane with a depth of 500 mm and a radius of 250 mm. A pressure head boundary condition of 5.7
289 mm was imposed on the soil surface delimited by the ring, while free drainage was set at the bottom
290 of the modeled domain. The values of θ_r , θ_s , η , and h_b obtained with the evaporation method were
291 used as initial values, to improve the fitting results. Note that the Brooks and Corey models were
292 considered for the water retention and hydraulic conductivity functions, in accordance with the four
293 analytical approaches. The water content parameters θ_r and θ_s were kept fixed and the tortuosity
294 parameter, l , was set to 0.5. Through a least-squares inverse solution routine, η , h_b and $K_{fs-HYDRUS}$
295 values were optimized using the measured cumulative infiltration data. **Table 3** summarizes the
296 optimized parameters, and an example for each soil of the inverse modeling is depicted in **Figure 2**.
297 For the four sampled soils, $K_{fs-HYDRUS}$ ranged from 28.2 to 839.9 mm h⁻¹. The wide range of K_{fs-}
298 $HYDRUS$ values supported the choice to test the proposed model, and the 14 different scenarios, on
299 these four hydraulically distinct soils.

300 **3.6. Data analysis**

301 In this investigation, we considered a total of 14 different scenarios to estimate K_{fs} data. More
302 specifically, the K_{fs} values were estimated by:

- 303 • **Approach 1 (scenario i):** determining λ through Eq. (6) and θ_i and θ_s from sampled soil
304 cores, and then fitting Eq. (1) to cumulative infiltration;
- 305 • **Approach 2 (scenarios ii-v):** determining λ , θ_i and θ_s , and introducing the three datasets
306 of c_2 and c_1 values, obtained with the CI, CL and DL fitting methods, into Eqs. (7) and
307 (8), and the c_4 and c_3 values into Eqs. (9) and (10);
- 308 • **Approach 3 (scenarios vi-ix):** estimating λ through Eq. (6) and introducing the three
309 datasets of c_2 estimates into Eq. (11), and the c_4 values into Eq. (12);

- 310 • **Approach 4 (scenarios x-xiii)**: using $\lambda = 150$ mm and introducing the three datasets of c_2
 311 estimates into Eq. (11), and the c_4 values into Eq. (12);
- 312 • **SSBI method (scenario xiv)** using $\lambda = 150$ mm and introducing the c_4 values into Eq.
 313 (13).

314 With reference to Approaches 1 and 3, λ values obtained from water retention data and estimated
 315 by Eq. (6) were averaged to obtain four site-representative values. A single value of θ_i and θ_s was
 316 also obtained for a given site by averaging individual determinations (Approach 1 and 2).

317 The field-saturated soil hydraulic conductivity, K_{fs} , estimates were compared with the
 318 corresponding values obtained by inverse solution from HYDRUS-2D/3D (i.e., the $K_{fs-HYDRUS}$
 319 values) using the relative error, $Er(K_{fs})$, defined as follows:

$$320 \quad Er(K_{fs}) = 100 \times \frac{K_{fs} - K_{fs-HYDRUS}}{K_{fs-HYDRUS}} \quad (14)$$

321 Note that positive $Er(K_{fs})$ values indicate overestimations, whereas negative values indicate
 322 underestimation. Small deviations, i.e., $Er(K_{fs}) \sim 0$, suggest that the estimates are close to actual
 323 values. $Er(K_{fs})$ values between -50% and +100% represent a factor of difference $f_D < 2$ between
 324 estimated and actual values. $Er(K_{fs})$ values between -66.7% and +200% represent $f_D < 3$. The factor
 325 of difference can be calculated as the ratio between the maximum and minimum of K_{fs} and the
 326 corresponding $K_{fs-HYDRUS}$ value [i.e., $f_D = MAX(K_{fs}, K_{fs-HYDRUS})/MIN(K_{fs}, K_{fs-HYDRUS})$]. Following
 327 Elrick and Reynolds (1992), f_D values not exceeding a value of two were considered indicative of
 328 similar estimates. Also note that all of the estimation and comparison procedures are synthesized in
 329 **Figure 2**.

330 For comparisons between paired observations, the paired differences, i.e., $K_{fs} - K_{fs-HYDRUS}$ for
 331 given scenario, were calculated and the hypothesis of normality of these differences was checked by
 332 the Kolmogorov-Smirnov test. For normally distributed data, a paired t -test was used to test the
 333 mean difference between paired observations at $P < 0.05$. For non-normally distributed data the

334 Wilcoxon signed rank test was used to test the median difference between paired observations at P
335 < 0.05 .

336 The adequacy of model fits was evaluated by checking the relative error, Er , and the root mean
337 squared differences, $RMSD$, defined as:

$$338 \quad Er = 100 \times \sqrt{\frac{\sum_{i=1}^n (x_i^{obs} - x_i)^2}{\sum_{i=1}^n (x_i^{obs})^2}} \quad (15)$$

$$339 \quad RMSD = \sqrt{\frac{\sum_{i=1}^n (x_i^{obs} - x_i)^2}{n}} \quad (16)$$

340 where n is the total number of data pairs, x_i^{obs} are the observed data and x_i are the values predicted
341 by the models. Values of $Er < 5\%$ were assumed indicative of a satisfactory fitting ability of the
342 models (Angulo-Jaramillo et al., 2016; Lassabatere et al., 2006).

343 **4. Results**

344 In the following subsections we present: i) the results of the analysis of single-ring infiltration
345 data and the performances of the different fitting methods (section 4.1), ii) the result of the
346 comparison between K_{fs} data estimated from different scenarios and those values obtained by
347 numerical inverse modeling with HYDRUS-2D/3D (section 4.2), and iii) a check for data reliability
348 by comparing K_{fs} estimates with laboratory measurements of unsaturated hydraulic conductivity, K_h
349 (section 4.3).

350 **4.1. Analysis of single-ring infiltration data**

351 We firstly used Eq. (4) to determine the time to steady-state, t_s , with the condition $\hat{E} > 2$
352 (**Figure 1a**). This threshold split the experimental data into two subsets that were then fitted to the
353 transient- ($t < t_s$) and steady-state ($t \geq t_s$) models. Time to steady-state ranged from 1.5–31.1 min,
354 depending on the run (**Table 4**). For the sandy soil, t_s was on average 19.1 min, with an infiltrated

355 depth $I(t_s)$ of 123.4 mm. In comparison, the sandy loam soils had mean t_s values of 5.2 minutes
356 (sandy-loam 1) and 2.2 minutes (sandy-loam 2). The sandy soil thus required a factor of 4 to 9 more
357 time to reach steady-state conditions compared to the sandy-loam soils, likely due to the presence of
358 a crust layer, which reduced infiltration rates (Alagna et al., 2019, 2013) and affected estimates for
359 infiltration parameters (Di Prima et al., 2018a).

360 During data analysis, peculiarities emerged within some of the infiltration datasets, with three
361 types of abnormal behaviors identified (**Figure 4**). In some runs, the early infiltration rates were
362 particularly high in comparison with the rest of the run (**Figure 4a**), causing a large initial jump in
363 cumulative infiltration (**Figure 4b, white circles**). This circumstance is quite common in coarse or
364 initially dry soils (Di Prima et al., 2016). In this case, the first data point of the I/\sqrt{t} vs. \sqrt{t} plot (CL
365 method) deviated from the general linear behavior (**Figure 4c, white circles**). This problem can be
366 easily solved by excluding the first data point from the cumulative infiltration (**Figure 4b**), allowing
367 the detection of a linear relationship (**Figure 4c, grey circles**), and a proper estimation of the c_1 and
368 c_2 coefficients. Such an adjustment was made on 14 infiltration runs, i.e., 35% of the cases. Other
369 investigations also suggested removing the early stage of the infiltration process when a
370 perturbation occurs (e.g., Bagarello et al., 2014c; Di Prima et al., 2018b; Vandervaere et al., 2000b).

371 One infiltration experiment, from the sandy soil, showed a sudden decrease in infiltration rate
372 (**Figure 4d**). This condition was not easily detectable from the visual inspection of the cumulative
373 infiltration curve (**Figure 4e**), but appeared when the data were linearized (**Figure 4f**). The lack of
374 linear data meant that Eq. (5a) was inappropriate and that the fitted parameters were physically
375 meaningless (Vandervaere et al., 2000a). For this reason the sample was excluded from subsequent
376 analyses. Possible contributing factors include water infiltrating into a less permeable layer (Alagna
377 et al., 2016; Lassabatere et al., 2019), air entrapment, vertical soil water content gradients and soil
378 sealing at the surface from repeated water applications (Bagarello et al., 2014c; Di Prima et al.,
379 2018a).

380 Other experiments had infiltration rates that increased with time (**Figure 4g**), such that the
381 cumulative infiltration curves exhibited convex shapes (**Figure 4h**). The fitting procedures applied
382 to these data produced negative values for the c_1 infiltration coefficient (**Figure 4i**). These cases
383 occurred at the sandy-loam 1 site (one instance) and sandy-loam 2 site (three instances), and reflect
384 that the early wetting phase was impeded due to hydrophobic surface films on soil particles and
385 non-zero contact angles between water and soil particles (Hallett et al., 2001; Jarvis et al., 2008).
386 Hydrophobia may be attributed to locally high OC content (Goebel et al., 2011) and exudates
387 produced by the plant root systems or living organisms like arbuscular mycorrhizal fungi (Rillig et
388 al., 2010). This effect is known to diminish during the wetting process (Alagna et al., 2018). It
389 should be noted that despite the water repellency, relatively high early infiltration rates were still
390 measured. This result indicates that the soils likely had subcritical water repellency (e.g., Di Prima
391 et al., 2017a; Lassabatere et al., 2019; Lichner et al., 2007; Lozano-Baez et al., 2018).

392 Three different sets of c_1 and c_2 values were obtained for transient-state data using the CI, CL
393 and DL methods. Overall, the c_1 and c_2 coefficients were properly estimated in 93% of the cases (37
394 of 40 runs) for the CI and CL methods, and 95% (38 of 40) for the DL method. The c_1 coefficient
395 ranged between 0.4 and 514.3 mm h^{-0.5} and the c_2 coefficient between 95.8 and 4424.2 mm h⁻¹
396 (**Table 5**). Differences between methods were more pronounced for c_1 compared to c_2 values, with
397 the latter only presenting statistically different estimates between the three procedures for the silty-
398 loam soil (**Figure 5**). Mean c_1 values were ordered as DL > CI > CL. Good fits (i.e., $Er < 5\%$) were
399 obtained for all cases except the DL method on the sandy-loam 2 site (mean $Er = 7.6\%$).

400 Finally, the analysis of steady-state data (i.e., the data points for which $\hat{E} \leq 2$; Eq. (4)) did not
401 show any such peculiarities, thus, the intercept, c_3 , and the slope, c_4 , of the regression line fitted to
402 the data points describing steady-state conditions could be properly estimated in all cases (**Table 5**).

403 4.2. Validation with HYDRUS predicted data

404 The inverse option in HYDRUS-2D/3D was used to optimize the η , h_b and $K_{fs-HYDRUS}$ parameters
405 on the measured cumulative infiltration data. The field-saturated soil hydraulic conductivity, K_{fs} ,
406 estimates obtained from the 14 different scenarios were compared with the corresponding values
407 obtained from HYDRUS, i.e., the $K_{fs-HYDRUS}$ values. For Approach 1, site-representative values of λ
408 (**Table 6**), θ_i and θ_s were considered, and K_{fs} was optimized fitting Eq. (1) to cumulative
409 infiltrations. The λ values were obtained by averaging for each soil the individual determination
410 obtained from Eq. (6) and considering the η and h_b parameters optimized on the retention data
411 obtained by the evaporation experiments. The K_{fs} values ranged between 29.2 and 429.1 mm h⁻¹
412 (**Table 7**), with 45 and 55% of the runs yielding respectively lower and higher K_{fs} estimates than the
413 HYDRUS-estimated values (**Figure 6**). The differences between K_{fs} and $K_{fs-HYDRUS}$ were non-
414 normally distributed according to the Kolmogorov-Smirnov test. The Wilcoxon signed rank test
415 showed that Approach 1 yielded K_{fs} estimates not significantly different from the $K_{fs-HYDRUS}$ values
416 (**Figures 7 and 8**). The relative error, $Er(K_{fs})$, ranged from -66.5 to 347.3%, with mean and median
417 factor of difference, f_D , values equal to 1.45 and 1.22. Individual values f_D were less than two in
418 85.0% and less than three in 97.5% of the cases, with only one case out of 40 yielding $f_D > 3$
419 (**Figure 9**). Therefore, K_{fs} estimates were acceptable in almost all cases when Eq. (1) was directly
420 fitted to experimental data.

421 For Approach 2, four sets of K_{fs} and λ values were determined: three sets for transient infiltration
422 data by Eqs. (12) and (13) (one set for each fitting procedure, i.e., CI, CL and DL), and one set for
423 steady-state data by Eqs. (14) and (15). The three transient scenarios yielded significant higher K_{fs}
424 estimates than the $K_{fs-HYDRUS}$ values (**Figures 7 and 8**), with mean f_D values equal to 12.23 (CI),
425 16.05 (CL), and 9.30 (DL), and individual f_D values higher than three in 80.0, 92.5 and 72.5% of the
426 cases (**Figure 9**). The steady-state scenario gave negative λ values, and consequentially negative
427 K_{fs} , in 75% of the cases (i.e., 30 out of 40). Overall, Approach 2 either poorly predicted λ and K_{fs}

428 data or failed to give valid estimates at all. The results obtained here can be viewed as a
429 confirmation of the conclusion by Stewart and Abou Najm (2018b).

430 For Approach 3, site-representative values of λ were calculated based on water retention
431 characteristics (**Table 6**) and four sets of K_{fs} values were determined: three sets for transient
432 infiltration data using Eq. (11), and one set for steady-state data using Eq. (12). For these scenarios,
433 K_{fs} ranged between 23.2 and 687.4 mm h⁻¹ (**Table 7**), with the transient scenarios yielding slightly
434 but significantly higher K_{fs} estimates than HYDRUS, and with the steady-state scenarios yielding
435 slightly but significantly lower estimates than HYDRUS (**Figures 7 and 8**). For the four scenarios,
436 f_D values were less than two in at least 75.0% of the transient scenarios and in 95.0% of the steady-
437 state cases (**Figure 9**), with mean f_D values ranging from 1.51 to 1.86 and median f_D values from
438 1.37 to 1.59. Given that K_{fs} estimates were acceptable in all cases, we considered a new scenario
439 (**Figures 6, 7, 8 and 9**) by averaging, for a given run, the four K_{fs} estimates. This newly conceived
440 scenario yielded lower mean and median f_D values, respectively equal to 1.46 and 1.24.

441 For Approach 4, a λ value of 150 mm was used to determine four sets of K_{fs} values (similar to
442 Approach 3). For the transient scenarios, using a λ value of 150 mm resulted in higher K_{fs} values
443 than were predicted by HYDRUS (**Figures 7 and 8**). For these three scenarios, f_D values were
444 higher than two in 67.5-82.5% of the cases and higher than three in 30.0-55.0% of the cases (**Figure**
445 **9**). Mean f_D values ranged from 2.71 to 3.52 and median f_D values varied from 2.59 to 3.27. Better
446 K_{fs} predictions were obtained by the steady-state scenario, with $Er(K_{fs})$ values ranging from -38.6 to
447 460.4%. This scenario yielded slightly but significant higher K_{fs} estimates than the actual values
448 (**Figures 7 and 8**), with mean and median f_D values equal to 1.57 and 1.49. Individual f_D values
449 were respectively less than two and three in 95.0 and 97.5% of the cases, and with only one case out
450 of 40 yielding $f_D > 3$ (**Figure 9**).

451 With the SSBI method, the $Er(K_{fs})$ values ranged from -36.8 to 476.6%, with mean and median
452 f_D values equal to 1.63 and 1.54. Individual values f_D were less than two in 90.0% and less than
453 three in 95.0% of the cases, with only two cases out of 40 yielding $f_D > 3$ (**Figure 9**).

454 4.3. Unsaturated vs. field-saturated soil hydraulic conductivity

455 A further check for data reliability was carried out by comparing K_{fs} estimates with laboratory
456 measurements of unsaturated hydraulic conductivity, K_h . **Table 8** summarizes the K_h values
457 measured by the unit hydraulic gradient laboratory method for the four soils. As expected, K_h
458 increased dramatically in the proximity of the saturation, i.e., for lower $|h_0|$ values (**Figure 10**). The
459 mean values of the unsaturated soil hydraulic conductivity obtained at the pressure head $h_0 = -10$
460 mm (K_{10}) ranged between 21.0 and 154.7 mm h⁻¹, with higher K_{10} values measured on the sandy
461 soil cores.

462 Firstly, it should be noted that, for the sandy soil, this comparison needs specific consideration
463 owing to crusting phenomena at the soil surface. In particular, the soil core collection process
464 disturbed the crust layer, such that fractures on the soil surface were observed in all soil cores.
465 Therefore, the upper layer of the soil was carefully removed in the laboratory and the measurements
466 of unsaturated hydraulic conductivity were conducted on the underlying, non-crust, soil layer. On
467 the contrary, in the field, we maintained the crust layer during the ponding experiments, in order to
468 give an insight on the potential of the applied model when a layered medium is characterized. As
469 mentioned above, the small insertion depth (i.e., 10 mm) of the ring used to run the Beerkan
470 experiments avoided the formation of fractures in the crust layer, ensuring that the measured
471 infiltration rates were indicative of the crust layer. As a consequence, 9 of 14 scenarios for this soil
472 produced mean values of K_{fs} lower than K_{10} , proving that the soil crust layer reduced water flow
473 during ponding experiments in the field.

474 For the silty-loam, sandy-loam 1 and sandy-loam 2 soils, K_{fs} determined from the 14 different
475 scenarios always remained higher than the measured K_h values. Therefore, physically possible K_{fs}
476 estimates were obtained in all cases, given that $K_{fs} > K_{10}$. For these soils, the 14 scenarios yielded
477 mean K_{fs} values that were 1.7–68.6 times higher than the corresponding K_{10} , i.e., up to two orders of
478 magnitude. Differences of this order of magnitude or even higher between saturated and near-

479 saturated hydraulic conductivity have been often observed under field conditions (e.g., Buczko et
480 al., 2003; Castellini et al., 2015; Di Prima et al., 2017a; Dunn and Phillips, 1991; Watson and
481 Luxmoore, 1986).

482 5. Discussion

483 The analysis of the cumulative infiltration measurements identified some runs with peculiarities
484 such as very high initial infiltration rates, undetectable linear relationships in the CL and DL
485 methods, and negative values of the infiltration coefficients. Still, infiltration data could be analyzed
486 to determine the constants c_1 and c_2 for 93% of the runs using the cumulative infiltration (CI) and
487 cumulative linearization (CL) methods, and for 95% of the runs using the differential linearization
488 (DL) method.

489 The infiltration constants were next applied to estimate K_{fs} using the comprehensive single-ring
490 infiltration model of Stewart and Abou Najm (2018a). Here, we considered four approaches and
491 thirteen scenarios that differed in how λ was constrained, while also comparing K_{fs} estimates using
492 the SBBI method. Approaches 1 and 3 were the most data demanding, requiring that λ was
493 estimated from water retention data and that soil samples were collected before the infiltration runs
494 to determine initial and saturated volumetric soil water contents (θ_i and θ_s), yet our analysis of the
495 field data showed that those approaches provided the most accurate K_{fs} estimates compared to
496 values obtained through numerical inverse modeling with HYDRUS-2D/3D. Approach 1 was the
497 most accurate overall, likely because it did not require any transformation of the infiltration data.
498 This approach is therefore recommended for situations when λ , θ_i and θ_s are well constrained. Still,
499 by averaging together the four K_{fs} estimates obtained by Approach 3 for a given run (i.e.,
500 considering together the scenarios vi – ix in **Figure 3**), the measurement uncertainty of that
501 approach was reduced to a level comparable to Approach 1. These averaged K_{fs} values avoided
502 uncertainties that might exist within each of the specific fitting procedures (CI, CL and DL), while
503 also overcoming any failed analyses (e.g., negative estimates for K_{fs}). As a result, this newly

504 considered scenario increased the accuracy of K_{fs} estimates, and as a result we recommend users
505 apply a similar averaging scheme when using Approach 3.

506 The K_{fs} estimates were less accurate but still acceptable for the steady-state scenarios of
507 Approach 4 and the SSBI method. The steady-state data likely provided better accuracy than the
508 transient data because the steady phase of the infiltration process avoids uncertainties due to
509 variations in infiltration rates caused by, for instance, soil sealing (Di Prima et al., 2018a) or water
510 repellency (Lichner et al., 2013). With both of these methods, no additional data are required to
511 determine K_{fs} , making these procedures desirable when surveying remote or large areas (Bagarello
512 et al., 2013). One difference between the two is that the SSBI method is theoretically usable for a
513 ponded depth of water on the infiltration surface, h_{source} , equal to zero and a null depth of ring
514 insertion into the soil, d (Bagarello et al., 2017), whereas both zero and positive values of both
515 h_{source} and d can be considered with Approach 4. Here both methods were used to analyze
516 infiltration runs that had a quasi-zero head of water imposed on the soil surface (Beerkan runs), so
517 the models performed similarly to one another.

518 The predictive potential of the model was also checked via comparison with laboratory
519 measurements of unsaturated hydraulic conductivity (K_h). For the silty-loam, sandy-loam 1 and
520 sandy-loam 2 soils, K_{fs} estimates from the 14 different scenarios were always higher than the
521 unsaturated soil hydraulic conductivity. Therefore, physically plausible K_{fs} values were obtained in
522 all cases. For the crusted sandy soil, $K_{fs} < K_h$ situations suggested that the surface crust layer
523 reduced water flow during ponding experiments in the field. In the future, measuring K_h values
524 directly in the field using a tension infiltrometer (Casey and Derby, 2002), or the portable Mini Disk
525 device (Decagon, 2014), may help to properly characterize unsaturated flow in crusted soils.
526 Indeed, field measurements are known to minimize soil disturbance in comparison with laboratory
527 methods performed on collected soil samples (Haverkamp et al., 1999). Moreover, tension
528 infiltrometers were successfully used in many investigations to characterize layered soils in the field
529 (e.g., Alagna et al., 2013; Di Prima et al., 2017b; Šimůnek et al., 1998; Vandervaere et al., 1997).

530 Altogether, the new comprehensive model and the underlying approaches to analyze single-ring
531 data may allow researchers to better approach heterogeneous datasets, including transient or steady-
532 state infiltration data and experiments carried out with different setups. The versatility of the new
533 model makes it a good candidate to successfully analyze the SWIG database developed by Rahmati
534 et al. (2018), which include 5023 infiltration curves collected across the world.

535 **6. Summary and conclusions**

536 In this study, we tested a new comprehensive model for single-ring data on four soils with
537 different textures, i.e., sandy, silty-loam and sandy-loam. The field-saturated soil hydraulic
538 conductivity, K_{fs} , values were estimated by four different approaches, which differ by the way they
539 derive K_{fs} , and constrain λ , θ_i and θ_s . For comparative purposes, the SSBI method was also applied
540 to estimate K_{fs} . In this investigation, we considered a total of 14 different scenarios to estimate K_{fs}
541 data that differed in the considered approach (i.e., Approaches 1-4 or SSBI), in the use of transient
542 or steady-state data, and in the fitting methods applied to transient data (CI, CL and DL).

543 The K_{fs} data estimated from different scenarios were compared for validation purposes with
544 those values obtained by numerical inverse modeling with HYDRUS-2D/3D. Among the different
545 scenarios, Approaches 1 and 3 appear as the more promising, yielding better K_{fs} predictions.
546 Conversely, the steady-state scenario of Approach 4 and the SSBI method are preferable when a
547 simplified experimental procedure is required, such as when sampling remote or large areas, given
548 that these interpretations do not require additional data and still provide acceptable estimates of K_{fs} .

549 **Acknowledgements**

550 This work was supported through the INFILTRON Project (ANR-17-CE04-0010) funded by the
551 French National Research Agency (ANR). S. Di Prima outlined the investigation and analyzed the
552 data. All authors contributed to discussing the results and writing the manuscript.

- 554 Alagna, V., Bagarello, V., Di Prima, S., Giordano, G., Iovino, M., 2016. Testing infiltration run effects on
 555 the estimated water transmission properties of a sandy-loam soil. *Geoderma* 267, 24–33.
 556 <https://doi.org/10.1016/j.geoderma.2015.12.029>
- 557 Alagna, V., Bagarello, V., Di Prima, S., Giordano, G., Iovino, M., 2013. A simple field method to measure
 558 the hydrodynamic properties of soil surface crust. *Journal of Agricultural Engineering* 44, 74–79.
 559 [https://doi.org/10.4081/jae.2013.\(s1\):e14](https://doi.org/10.4081/jae.2013.(s1):e14)
- 560 Alagna, V., Bagarello, V., Di Prima, S., Guitoli, F., Iovino, M., Keesstra, S., Cerdà, A., 2019. Using
 561 beerkan experiments to estimate hydraulic conductivity of a crusted loamy soil in a Mediterranean
 562 vineyard. *Journal of Hydrology and Hydromechanics* 67. <https://doi.org/10.2478/johh-2018-0023>
- 563 Alagna, V., Iovino, M., Bagarello, V., Mataix-Solera, J., Lichner, L., 2018. Alternative analysis of transient
 564 infiltration experiment to estimate soil water repellency. *Hydrological Processes*.
 565 <https://doi.org/10.1002/hyp.13352>
- 566 Angulo-Jaramillo, R., Bagarello, V., Iovino, M., Lassabatere, L., 2016. Saturated Soil Hydraulic
 567 Conductivity, in: *Infiltration Measurements for Soil Hydraulic Characterization*. Springer
 568 International Publishing, pp. 43–180. https://doi.org/10.1007/978-3-319-31788-5_2
- 569 Angulo-Jaramillo, R., Vandervaere, J.-P., Roulier, S., Thony, J.-L., Gaudet, J.-P., Vauclin, M., 2000. Field
 570 measurement of soil surface hydraulic properties by disc and ring infiltrometers: A review and recent
 571 developments. *Soil and Tillage Research* 55, 1–29. [https://doi.org/10.1016/S0167-1987\(00\)00098-2](https://doi.org/10.1016/S0167-1987(00)00098-2)
- 572 Bagarello, V., Castellini, M., Di Prima, S., Giordano, G., Iovino, M., 2013. Testing a Simplified Approach to
 573 Determine Field Saturated Soil Hydraulic Conductivity. *Procedia Environmental Sciences* 19, 599–
 574 608. <https://doi.org/10.1016/j.proenv.2013.06.068>
- 575 Bagarello, V., Castellini, M., Iovino, M., 2007. Comparison of unconfined and confined unsaturated
 576 hydraulic conductivity. *Geoderma* 137, 394–400. <https://doi.org/10.1016/j.geoderma.2006.08.031>
- 577 Bagarello, V., Di Prima, S., Giordano, G., Iovino, M., 2014a. A test of the Beerkan Estimation of Soil
 578 Transfer parameters (BEST) procedure. *Geoderma* 221–222, 20–27.
 579 <https://doi.org/10.1016/j.geoderma.2014.01.017>
- 580 Bagarello, V., Di Prima, S., Iovino, M., 2017. Estimating saturated soil hydraulic conductivity by the near
 581 steady-state phase of a Beerkan infiltration test. *Geoderma* 303, 70–77.
 582 <https://doi.org/10.1016/j.geoderma.2017.04.030>
- 583 Bagarello, V., Di Prima, S., Iovino, M., 2014b. Comparing Alternative Algorithms to Analyze the Beerkan
 584 Infiltration Experiment. *Soil Science Society of America Journal* 78, 724.
 585 <https://doi.org/10.2136/sssaj2013.06.0231>
- 586 Bagarello, V., Di Prima, S., Iovino, M., Provenzano, G., 2014c. Estimating field-saturated soil hydraulic
 587 conductivity by a simplified Beerkan infiltration experiment. *Hydrological Processes* 28, 1095–1103.
 588 <https://doi.org/10.1002/hyp.9649>
- 589 Bagarello, V., Iovino, M., Reynolds, W., 1999. Measuring hydraulic conductivity in a cracking clay soil
 590 using the Guelph permeameter. *Transactions of the ASAE* 42.
- 591 Braud, I., De Condappa, D., Soria, J.M., Haverkamp, R., Angulo-Jaramillo, R., Galle, S., Vauclin, M., 2005.
 592 Use of scaled forms of the infiltration equation for the estimation of unsaturated soil hydraulic
 593 properties (the Beerkan method). *European Journal of Soil Science* 56, 361–374.
 594 <https://doi.org/10.1111/j.1365-2389.2004.00660.x>
- 595 Braud, I., Desprats, J.-F., Ayrat, P.-A., Bouvier, C., Vandervaere, J.-P., 2017. Mapping topsoil field-
 596 saturated hydraulic conductivity from point measurements using different methods. *Journal of*
 597 *Hydrology and Hydromechanics* 65. <https://doi.org/10.1515/johh-2017-0017>
- 598 Brooks, R.H., Corey, T., 1964. hydraulic properties of porous media. *Hydrol. Paper 3.*, Colorado State
 599 University, Fort Collins.
- 600 Buczko, U., Benz, O., Hangen, E., Brunotte, J., Huttl, R., 2003. Infiltration and macroporosity of a silt loam
 601 soil under two contrasting tillage systems. *Landbauforschung Volkenrode* 53, 181–190.
- 602 Casey, F.X.M., Derby, N.E., 2002. Improved design for an automated tension infiltrometer. *Soil Sci. Soc.*
 603 *Am. J.* 66, 64–67.

604 Castellini, M., Di Prima, S., Iovino, M., 2018. An assessment of the BEST procedure to estimate the soil
605 water retention curve: A comparison with the evaporation method. *Geoderma* 320, 82–94.
606 <https://doi.org/10.1016/j.geoderma.2018.01.014>

607 Castellini, M., Giglio, L., Niedda, M., Palumbo, A.D., Ventrella, D., 2015. Impact of biochar addition on the
608 physical and hydraulic properties of a clay soil. *Soil and Tillage Research* 154, 1–13.
609 <https://doi.org/10.1016/j.still.2015.06.016>

610 Decagon, 2014. *Minidisk Infiltrometer User’s Manual*. Decagon Devices, Inc., Pullman, USA 24.

611 Di Prima, S., Bagarello, V., Angulo-Jaramillo, R., Bautista, I., Cerdà, A., del, C.A., González-Sanchis, M.,
612 Iovino, M., Lassabatere, L., Maetzke, F., 2017a. Impacts of thinning of a Mediterranean oak forest
613 on soil properties influencing water infiltration. *Journal of Hydrology and Hydromechanics* 65, 276–
614 286. <https://doi.org/10.1515/johh-2017-0016>

615 Di Prima, S., Bagarello, V., Lassabatere, L., Angulo-Jaramillo, R., Bautista, I., Burguet, M., Cerdà, A.,
616 Iovino, M., Prosdocimi, M., 2017b. Comparing Beerkan infiltration tests with rainfall simulation
617 experiments for hydraulic characterization of a sandy-loam soil. *Hydrological Processes* 31, 3520–
618 3532. <https://doi.org/10.1002/hyp.11273>

619 Di Prima, S., Concialdi, P., Lassabatere, L., Angulo-Jaramillo, R., Pirastru, M., Cerdà, A., Keesstra, S.,
620 2018a. Laboratory testing of Beerkan infiltration experiments for assessing the role of soil sealing on
621 water infiltration. *CATENA* 167, 373–384. <https://doi.org/10.1016/j.catena.2018.05.013>

622 Di Prima, S., Lassabatere, L., Bagarello, V., Iovino, M., Angulo-Jaramillo, R., 2016. Testing a new
623 automated single ring infiltrometer for Beerkan infiltration experiments. *Geoderma* 262, 20–34.
624 <https://doi.org/10.1016/j.geoderma.2015.08.006>

625 Di Prima, S., Lassabatere, L., Rodrigo-Comino, J., Marrosu, R., Pulido, M., Angulo-Jaramillo, R., Úbeda,
626 X., Keesstra, S., Cerdà, A., Pirastru, M., 2018b. Comparing Transient and Steady-State Analysis of
627 Single-Ring Infiltrometer Data for an Abandoned Field Affected by Fire in Eastern Spain. *Water* 10.
628 <https://doi.org/10.3390/w10040514>

629 Di Prima, S., Marrosu, R., Lassabatere, L., Angulo-Jaramillo, R., Pirastru, M., 2018c. In situ characterization
630 of preferential flow by combining plot- and point-scale infiltration experiments on a hillslope.
631 *Journal of Hydrology* 563, 633–642. <https://doi.org/10.1016/j.jhydrol.2018.06.033>

632 Di Prima, S., Rodrigo-Comino, J., Novara, A., Iovino, M., Pirastru, M., Keesstra, S., Cerdà, A., 2018d. Soil
633 Physical Quality of Citrus Orchards Under Tillage, Herbicide, and Organic Managements.
634 *Pedosphere* 28, 463–477. [https://doi.org/10.1016/S1002-0160\(18\)60025-6](https://doi.org/10.1016/S1002-0160(18)60025-6)

635 Dunn, G.H., Phillips, R.E., 1991. Macroporosity of a Well-Drained Soil under No-Till and Conventional
636 Tillage. *Soil Science Society of America Journal* 55, 817.
637 <https://doi.org/10.2136/sssaj1991.03615995005500030031x>

638 Dusek, J., Vogel, T., Dohnal, M., Gerke, H.H., 2012. Combining dual-continuum approach with diffusion
639 wave model to include a preferential flow component in hillslope scale modeling of shallow
640 subsurface runoff. *Advances in Water Resources* 44, 113–125.
641 <https://doi.org/10.1016/j.advwatres.2012.05.006>

642 Elrick, D.E., Reynolds, W.D., 1992. Methods for analyzing constant-head well permeameter data. *Soil*
643 *Science Society of America Journal* 56, 320.
644 <https://doi.org/10.2136/sssaj1992.03615995005600010052x>

645 Gee, G.W., Bauder, J.W., 1986. Particle-size Analysis, in: SSSA Book Series, Klute, A. (Ed.), *Methods of*
646 *Soil Analysis, Part 1: Physical and Mineralogical Methods*. Soil Science Society of America,
647 American Society of Agronomy, pp. 383–411.

648 Goebel, M.-O., Bachmann, J., Reichstein, M., Janssens, I.A., Guggenberger, G., 2011. Soil water repellency
649 and its implications for organic matter decomposition – is there a link to extreme climatic events?
650 *Global Change Biology* 17, 2640–2656. <https://doi.org/10.1111/j.1365-2486.2011.02414.x>

651 Hallett, P.D., Baumgartl, T., Young, I.M., 2001. Subcritical water repellency of aggregates from a range of
652 soil management practices. *Soil Science Society of America Journal* 65, 184–190.

653 Haverkamp, R., Bouraoui, F., Zammit, C., Angulo-Jaramillo, R., 1999. Soil properties and moisture
654 movement in the unsaturated zone. *Handbook of groundwater engineering*.

655 Jarvis, N., Etana, A., Stagnitti, F., 2008. Water repellency, near-saturated infiltration and preferential solute
656 transport in a macroporous clay soil. *Geoderma* 143, 223–230.
657 <https://doi.org/10.1016/j.geoderma.2007.11.015>

- 658 Klute, A., Dirksen, C., 1986. Hydraulic Conductivity and Diffusivity: Laboratory Methods. *Methods of Soil*
659 *Analysis: Part 1—Physical and Mineralogical Methods* sssabookseries, 687–734.
660 <https://doi.org/10.2136/sssabookser5.1.2ed.c28>
- 661 Lassabatere, L., Angulo-Jaramillo, R., Soria Ugalde, J.M., Cuenca, R., Braud, I., Haverkamp, R., 2006.
662 Beerkan estimation of soil transfer parameters through infiltration experiments—BEST. *Soil Science*
663 *Society of America Journal* 70, 521. <https://doi.org/10.2136/sssaj2005.0026>
- 664 Lassabatere, L., Di Prima, S., Angulo-Jaramillo, R., Keesstra, S., Salesa, D., 2019. Beerkan multi-runs for
665 characterizing water infiltration and spatial variability of soil hydraulic properties across scales.
666 *Hydrological Sciences Journal*.
- 667 Lichner, L., Hallett, P., Feeney, D., Ďugová, O., Šír, M., Tesař, M., 2007. Field measurement of soil water
668 repellency and its impact on water flow under different vegetation. *Biologia* 62.
669 <https://doi.org/10.2478/s11756-007-0106-4>
- 670 Lichner, L., Hallett, P.D., Drongová, Z., Czachor, H., Kovacik, L., Mataix-Solera, J., Homolák, M., 2013.
671 Algae influence the hydrophysical parameters of a sandy soil. *CATENA* 108, 58–68.
672 <https://doi.org/10.1016/j.catena.2012.02.016>
- 673 Lozano-Baez, S.E., Cooper, M., Ferraz, S.F.B., Ribeiro Rodrigues, R., Pirastru, M., Di Prima, S., 2018.
674 Previous Land Use Affects the Recovery of Soil Hydraulic Properties after Forest Restoration. *Water*
675 10. <https://doi.org/10.3390/w10040453>
- 676 Marquardt, D.W., 1963. An Algorithm for Least-Squares Estimation of Nonlinear Parameters. *Journal of the*
677 *Society for Industrial and Applied Mathematics* 11, 431–441.
- 678 Mertens, J., Jacques, D., Vanderborght, J., Feyen, J., 2002. Characterisation of the field-saturated hydraulic
679 conductivity on a hillslope: in situ single ring pressure infiltrometer measurements. *Journal of*
680 *Hydrology* 263, 217–229. [https://doi.org/10.1016/S0022-1694\(02\)00052-5](https://doi.org/10.1016/S0022-1694(02)00052-5)
- 681 Mubarak, I., Mailhol, J.C., Angulo-Jaramillo, R., Ruelle, P., Boivin, P., Khaleedian, M., 2009. Temporal
682 variability in soil hydraulic properties under drip irrigation. *Geoderma* 150, 158–165.
683 <https://doi.org/10.1016/j.geoderma.2009.01.022>
- 684 Philip, J., 1957. The theory of infiltration: 4. Sorptivity and algebraic infiltration equations. *Soil sci* 84, 257–
685 264.
- 686 Rahmati, M., Weihermüller, L., Vanderborght, J., Pachepsky, Y.A., Mao, L., Sadeghi, S.H., Moosavi, N.,
687 Kheirfam, H., Montzka, C., Looy, K.V., Toth, B., Hazbavi, Z., Yamani, W.A., Albalasmeh, A.A.,
688 Alghzawi, M.Z., Angulo-Jaramillo, R., Antonino, A.C.D., Arampatzis, G., Armindo, R.A., Asadi,
689 H., Bamutaze, Y., Batlle-Aguilar, J., Bechet, B., Becker, F., Blöschl, G., Bohne, K., Braud, I.,
690 Castellano, C., Cerdà, A., Chalhoub, M., Cichota, R., Císlarová, M., Clothier, B., Coquet, Y.,
691 Cornelis, W., Corradini, C., Coutinho, A.P., Oliveira, M.B. de, Macedo, J.R. de, Durães, M.F.,
692 Emami, H., Eskandari, I., Farajnia, A., Flammini, A., Fodor, N., Gharaibeh, M., Ghavimippanah,
693 M.H., Ghezzehei, T.A., Giertz, S., Hatzigiannakis, E.G., Horn, R., Jiménez, J.J., Jacques, D.,
694 Keesstra, S.D., Kelishadi, H., Kiani-Harchegani, M., Kouselou, M., Kumar Jha, M., Lassabatere, L.,
695 Li, X., Liebig, M.A., Lichner, L., López, M.V., Machiwal, D., Mallants, D., Mallmann, M.S.,
696 Marques, O., De, J.D., Marshall, M.R., Mertens, J., Meunier, F., Mohammadi, M.H., Mohanty, B.P.,
697 Moncada, M.P., Montenegro, S., Morbidelli, R., Moret-Fernández, D., Moosavi, A.A., Mosaddeghi,
698 M.R., Mousavi, S.B., Mozaffari, H., Nabiollahi, K., Neyshabouri, M.R., Ottoni, M.V., Filho, O.,
699 Benedicto, T., Rad, P., Reza, M., Panagopoulos, A., Peth, S., Peyneau, P.-E., Picciafuoco, T.,
700 Poesen, J., Pulido, M., Reinert, D.J., Reinsch, S., Rezaei, M., Roberts, F.P., Robinson, D., Rodrigo-
701 Comino, J., Filho, R., Corrêa, O., Saito, T., Suganuma, H., Saltalippi, C., Sándor, R., Schütt, B.,
702 Seeger, M., Sepehrnia, N., Sharifi Moghaddam, E., Shukla, M., Shutaro, S., Sorando, R., Stanley,
703 A.A., Strauss, P., Su, Z., Taghizadeh-Mehrjardi, R., Taguas, E., Teixeira, W.G., Vaezi, A.R.,
704 Vafakhah, M., Vogel, T., Vogeler, I., Votrubova, J., Werner, S., Winarski, T., Yilmaz, D., Young,
705 M.H., Zacharias, S., Zeng, Y., Zhao, Y., Zhao, H., Vereecken, H., 2018. Development and Analysis
706 of Soil Water Infiltration Global Database. *Earth System Science Data Discussions* 1–42.
707 <https://doi.org/10.5194/essd-2018-11>
- 708 Reynolds, W., Elrick, D., 1985. In situ measurement of field-saturated hydraulic conductivity, sorptivity, and
709 the α -parameter using the Guelph permeameter. *Soil Science* 140, 292–302.
- 710 Reynolds, W., Elrick, D., Youngs, E., 2002. 3.4.3.2.a. Single-ring and double- or concentric-ring
711 infiltrometers. *Methods of Soil Analysis, Part 4, Physical Methods*, J.H. Dane and G.C. Topp co-
712 editors, Number 5 in the Soil Science Society of America Book Series, Soil Science Society of
713 America, Inc. Madison, Wisconsin, USA, pp. 821–826.

- 714 Reynolds, W.D., Bowman, B.T., Brunke, R.R., Drury, C.F., Tan, C.S., 2000. Comparison of tension
715 infiltrometer, pressure infiltrometer, and soil core estimates of saturated hydraulic conductivity. *Soil*
716 *Science Society of America Journal* 64, 478–484. <https://doi.org/10.2136/sssaj2000.642478x>
- 717 Reynolds, W.D., Elrick, D.E., 1990. Poned Infiltration From a Single Ring: I. Analysis of Steady Flow. *Soil*
718 *Science Society of America Journal* 54, 1233.
719 <https://doi.org/10.2136/sssaj1990.03615995005400050006x>
- 720 Rillig, M.C., Mardatin, N.F., Leifheit, E.F., Antunes, P.M., 2010. Mycelium of arbuscular mycorrhizal fungi
721 increases soil water repellency and is sufficient to maintain water-stable soil aggregates. *Soil*
722 *Biology and Biochemistry* 42, 1189–1191. <https://doi.org/10.1016/j.soilbio.2010.03.027>
- 723 Sakaguchi, A., Nishimura, T., Kato, M., 2005. The Effect of Entrapped Air on the Quasi-Saturated Soil
724 Hydraulic Conductivity and Comparison with the Unsaturated Hydraulic Conductivity. *Vadose Zone*
725 *Journal* 4, 139–144. <https://doi.org/10.2136/vzj2005.0139>
- 726 Seki, K., 2007. SWRC fit - a nonlinear fitting program with a water retention curve for soils having
727 unimodal and bimodal pore structure. *Hydrology and Earth System Sciences Discussions* 4, 407–
728 437. <https://doi.org/10.5194/hessd-4-407-2007>
- 729 Šimůnek, J., Angulo-Jaramillo, R., Schaap, M.G., Vandervaere, J.-P., van Genuchten, M.T., 1998. Using an
730 inverse method to estimate the hydraulic properties of crusted soils from tension-disc infiltrometer
731 data. *Geoderma* 86, 61–81. [https://doi.org/10.1016/S0016-7061\(98\)00035-4](https://doi.org/10.1016/S0016-7061(98)00035-4)
- 732 Šimůnek, J., Hopmans, J.W., 2002. Parameter Optimization and Nonlinear Fitting. *SSSA Book Series,*
733 *Methods of Soil Analysis: Part 4 Physical Methods* 5.4, 139–157.
734 <https://doi.org/10.2136/sssabookser5.4.c7>
- 735 Šimůnek, J., van Genuchten, M.T., Šejna, M., 2008. Development and Applications of the HYDRUS and
736 STANMOD Software Packages and Related Codes. *Vadose Zone Journal* 7, 587.
737 <https://doi.org/10.2136/vzj2007.0077>
- 738 Smiles, D., Knight, J., 1976. A note on the use of the Philip infiltration equation. *Soil Res.* 14, 103–108.
- 739 Souza, E.S., Antonino, A.C.D., Heck, R.J., Montenegro, S.M.G.L., Lima, J.R.S., Sampaio, E.V.S.B.,
740 Angulo-Jaramillo, R., Vauclin, M., 2014. Effect of crusting on the physical and hydraulic properties
741 of a soil cropped with Castor beans (*Ricinus communis* L.) in the northeastern region of Brazil. *Soil*
742 *and Tillage Research* 141, 55–61. <https://doi.org/10.1016/j.still.2014.04.004>
- 743 Stewart, R.D., Abou Najm, M.R., 2018a. A Comprehensive Model for Single Ring Infiltration I: Initial
744 Water Content and Soil Hydraulic Properties. *Soil Science Society of America Journal* 0, 0.
745 <https://doi.org/10.2136/sssaj2017.09.0313>
- 746 Stewart, R.D., Abou Najm, M.R., 2018b. A Comprehensive Model for Single Ring Infiltration II: Estimating
747 Field-Saturated Hydraulic Conductivity. *Soil Science Society of America Journal* 0, 0.
748 <https://doi.org/10.2136/sssaj2017.09.0314>
- 749 Van Bemmelen, J., 1890. Über die Bestimmung des Wassers, des Humus, des Schwefels, der in den
750 colloidalen Silikaten gebundenen Kieselsäure, des Mangans usw im Ackerboden. *Die*
751 *Landwirtschaftlichen Versuchs-Stationen* 37, 279–290.
- 752 Vandervaere, J.-P., Peugeot, C., Vauclin, M., Angulo Jaramillo, R., Lebel, T., 1997. Estimating hydraulic
753 conductivity of crusted soils using disc infiltrometers and minitensiometers. *Journal of Hydrology,*
754 *HAPEX-Sahel* 188–189, 203–223. [https://doi.org/10.1016/S0022-1694\(96\)03160-5](https://doi.org/10.1016/S0022-1694(96)03160-5)
- 755 Vandervaere, J.-P., Vauclin, M., Elrick, D.E., 2000a. Transient flow from tension infiltrometers I. The two-
756 parameter equation. *Soil Science Society of America Journal* 64, 1263–1272.
- 757 Vandervaere, J.-P., Vauclin, M., Elrick, D.E., 2000b. Transient Flow from Tension Infiltrometers II. Four
758 Methods to Determine Sorptivity and Conductivity. *Soil Science Society of America Journal* 64,
759 1272–1284.
- 760 Verbist, K., Torfs, S., Cornelis, W.M., Oyarzún, R., Soto, G., Gabriels, D., 2010. Comparison of single- and
761 double-ring infiltrometer methods on stony soils. *Vadose Zone Journal* 9, 462–475.
762 <https://doi.org/10.2136/vzj2009.0058>
- 763 Walkley, A., Black, I.A., 1934. An examination of the Degtjareff method for determining soil organic
764 matter, and a proposed modification of the chromic acid titration method. *Soil science* 37, 29–38.
- 765 Watson, K.W., Luxmoore, R.J., 1986. Estimating macroporosity in a forest watershed by use of a tension
766 infiltrometer. *Soil Science Society of America Journal* 50, 578–582.
- 767 Wind, G.P., 1969. Capillary conductivity data estimated by a simple method. *Water In The Unsaturated Zone*
768 *Proc Wageningen Symp.*

769 Wu, L., Pan, L., 1997. A generalized solution to infiltration from single-ring infiltrometers by scaling. Soil
770 Science Society of America Journal 61, 1318–1322.
771 Wu, L., Pan, L., Mitchell, J., Sanden, B., 1999. Measuring Saturated Hydraulic Conductivity using a
772 Generalized Solution for Single-Ring Infiltrimeters. Soil Science Society of America Journal 63,
773 788. <https://doi.org/10.2136/sssaj1999.634788x>
774 Yilmaz, D., Lassabatere, L., Angulo-Jaramillo, R., Deneele, D., Legret, M., 2010. Hydrodynamic
775 Characterization of Basic Oxygen Furnace Slag through an Adapted BEST Method. Vadose Zone
776 Journal 9, 107. <https://doi.org/10.2136/vzj2009.0039>
777 Zhang, R., 1997. Determination of Soil Sorptivity and Hydraulic Conductivity from the Disk Infiltrimeter.
778 Soil Science Society of America Journal 61, 1024.
779 <https://doi.org/10.2136/sssaj1997.03615995006100040005x>
780

781

782 **Table 1.** Coordinates, soil textural classification, % clay (0–2 μm), % silt (2–50 μm), and % sand (50–2000
783 μm) content (size classes based on USDA classification system) in the 0-10 cm depth range, soil organic
784 matter content (SOM in %), initial volumetric soil water content (θ_i in $\text{cm}^3\text{cm}^{-3}$), and dry soil bulk
785 density (ρ_b in g cm^{-3}) for the four sampled soils (sample size for each soil, $N = 10$). Standard deviations
786 are indicated in parentheses.

Coordinates	39°46'51"N 8°33'12"E	41°27'4"N 15°30'4"E	38°6'25"N 13°21'6"E	38°4'53"N 13°25'7"E
Soil use	Corn	Durum wheat	Citrus orchard	Citrus orchard
Soil management	Tilled four months before with spreading of sewage (liquid cow manure)	Tilled six months before	Undisturbed	Undisturbed Tilled about two or three months before
Soil texture	Sandy	Silty-loam	Sandy-loam	Sandy-loam
Clay (%)	4.5 (2.2)	13.0 (1.7)	17.6 (1.9)	14.5 (3.3)
Silt (%)	5.0 (1.3)	60.7 (1.7)	29.8 (2.8)	22.7 (2.0)
Sand (%)	90.4 (2.1)	26.3 (2.3)	52.6 (4.7)	62.8 (1.8)
SOM (%)	1.8 (0.04)	2.7 (0.05)	3.9 (0.7)	2.0 (0.3)
θ_i ($\text{cm}^3 \text{cm}^{-3}$)	0.150 (0.03)	0.141 (0.02)	0.118 (0.01)	0.139 (0.02)
ρ_b (g cm^{-3})	1.198 (9.8)	1.128 (7.5)	1.127 (4.2)	1.315 (8.0)

787
788

789 **Table 2.** Mean values of the parameters obtained by fitting the Brooks and Corey model to the water
 790 retention data collected during the evaporation experiments. The coefficients of variation (%) are listed
 791 in parentheses.

Soil	Variable					
	θ_s	θ_r	h_b	η	<i>RMSD</i>	<i>Er</i>
Sandy	0.439 (11.4)	0.129 (70.2)	132.9 (26.3)	3.968 (38.7)	0.005 (29.5)	1.67 (26.1)
Silty-loam	0.491 (13.0)	0.130 (91.0)	248.4 (55.7)	3.045 (29.1)	0.003 (36.0)	0.76 (34.4)
Sandy-loam 1	0.365 (11.5)	0.025 (113.9)	227.8 (17.7)	2.822 (5.0)	0.003 (22.8)	1.18 (29.5)
Sandy-loam 2	0.458 (4.0)	0.126 (62.1)	276.4 (8.5)	3.122 (13.4)	0.004 (39.4)	1.19 (44.6)

792 θ_s = saturated volumetric soil water content determined based on the water content of the saturated cores
 793 ($\text{cm}^3 \text{cm}^{-3}$); θ_r = residual volumetric soil water content ($\text{cm}^3 \text{cm}^{-3}$); h_b = head scale parameter (mm); η =
 794 pores size index (-); *RMSD* = root mean squared differences ($\text{cm}^3 \text{cm}^{-3}$); *Er* = relative error (%).

795

796 **Table 3.** Mean values of the parameters obtained by inverse solution from HYDRUS-2D/3D. The
 797 coefficients of variation (%) are listed in parentheses.

Soil	Variable				
	$K_{fs-HYDRUS}$	h_b	η	$RMSD$	Er
Sandy	76.8 (44.7)	92.0 (51.5)	5.344 (30.5)	6.0 (61)	5.9 (61)
Silty-loam	47.3 (45.9)	157.6 (38.3)	4.421 (29.1)	1.1 (94.6)	1.0 (94.6)
Sandy-loam 1	130.2 (62.5)	165.3 (92)	6.381 (5.9)	3.8 (46.4)	3.6 (46.4)
Sandy-loam 2	460.1 (49.4)	104.7 (44.4)	6.181 (16.3)	5.1 (34.8)	5.0 (34.8)

798 $K_{fs-HYDRUS}$ = field-saturated soil hydraulic conductivity (mm h^{-1}); h_b = head scale parameter (mm); η = pores
 799 size index (-); $RMSD$ = root mean squared differences (mm); Er = relative error (%).

800

801 **Table 4.** Minimum (Min), maximum (Max), mean, median and coefficient of variation (CV, in %) of the
 802 equilibration time, t_s (min), and infiltrated depth at the equilibration time, $I(t_s)$ (mm). $N = 10$ samples for
 803 each soil.

Variable	Soil	Statistic				
		Min	Max	Mean	Median	CV
t_s	Sandy	12.8	31.1	19.1	17.1	35.0
	Silty-loam	11.6	24.8	17.0	16.0	24.9
	Sandy-loam 1	2.6	10.1	5.2	3.4	57.1
	Sandy-loam 2	1.5	3.4	2.2	1.9	29.2
$I(t_s)$	Sandy	113.2	135.8	123.4	124.5	6.8
	Silty-loam	101.9	124.5	121.1	124.5	6.3
	Sandy-loam 1	56.6	124.5	99.6	107.5	21.3
	Sandy-loam 2	101.9	135.8	125.6	124.5	9.0

804

805

806 **Table 5.** Sample size (N), minimum (Min), maximum (Max), mean, median and coefficient of variation
807 (CV, in %) of the c_1 (mm h^{-0.5}) and c_2 (mm h⁻¹) parameters estimated from transient infiltration data by
808 the cumulative infiltration (CI), cumulative linearization (CL) and differential linearization (DL)
809 methods, the relative errors, Er (%), of the fitting of the functional relationships to the experimental data,
810 and the intercept, c_3 (mm), and slope, c_4 (mm h⁻¹), of the regression line fitted to the last data points
811 describing the steady-state conditions on the I vs. t plot.

Soil	Type of data	Fitting method	Variable	Statistic						
				N	Min	Max	Mean	Median	CV	
Sandy	Transient data	CI	c_1	10	40.9	119.9	73.4	66.6	32.8	
			c_2	10	115.9	463.5	296.7	294.6	33.0	
			Er	10	0.3	4.9	1.4	1.2	98.1	
		CL	c_1	9	38.8	83.4	63.0	60.3	22.7	
			c_2	9	195.1	472.9	330.5	307.9	23.6	
			Er	9	0.5	2.6	1.4	1.4	57.7	
		DL	c_1	9	52.7	155.5	95.2	85.8	37.1	
			c_2	9	163.3	417.9	274.0	240.5	32.0	
			Er	9	1.4	12.8	4.7	3.9	76.0	
	Steady data		c_3	10	21.5	74.5	41.2	39.8	42.9	
		c_4	10	120.3	454.4	288.6	290.4	35.6		
Silty-loam	Transient data	CI	c_1	10	84.5	147.4	118.0	118.3	19.1	
			c_2	10	95.8	313.7	216.9	214.4	29.1	
			Er	10	0.3	1.8	1.1	1.0	44.5	
		CL	c_1	10	41.3	86.5	59.5	55.2	27.1	
			c_2	10	167.8	477.2	315.5	340.1	31.1	
			Er	10	0.7	3.1	1.8	1.8	43.0	
		DL	c_1	10	85.6	121.1	101.4	98.6	12.4	
			c_2	10	133.6	366.0	243.9	254.0	29.8	
			Er	10	1.4	5.7	2.6	2.2	46.9	
	Steady data		c_3	10	29.8	42.5	36.7	37.4	9.5	
			c_4	10	219.2	448.6	313.5	328.9	24.9	
	Sandy-loam 1	Transient data	CI	c_1	10	35.1	150.6	84.4	80.7	48.7
				c_2	10	245.6	2518.6	1194.4	1388.8	56.6
				Er	10	0.2	2.6	1.5	1.5	55.4
			CL	c_1	10	6.2	126.9	63.1	64.9	65.5
c_2				10	233.5	2674.2	1263.7	1399.3	55.7	
Er				10	0.2	3.7	1.6	1.1	76.6	
DL			c_1	9	5.1	309.7	114.3	106.1	93.4	
			c_2	9	303.9	1742.9	1161.3	1240.4	39.7	
			Er	9	0.2	9.4	4.2	4.4	63.0	
Steady data			c_3	10	7.1	38.1	19.4	19.8	45.4	
			c_4	10	289.5	2077.9	1192.4	1331.6	46.1	
Sandy-loam 2		Transient data	CI	c_1	7	53.0	251.2	135.3	135.2	55.6
				c_2	7	2090.3	4183.2	3192.3	3399.1	24.8
				Er	7	0.8	3.0	1.9	1.9	34.2
			CL	c_1	8	5.8	181.4	97.5	101.5	64.8
	c_2			8	2165.5	4424.2	3197.0	3236.1	26.8	
	Er			8	0.9	3.9	2.4	2.1	48.3	
	DL		c_1	10	0.4	514.3	247.7	236.6	65.0	
			c_2	10	1119.4	3727.1	2608.8	2742.0	32.6	
			Er	10	2.6	11.6	7.6	7.8	36.7	
	Steady data		c_3	10	5.3	54.1	28.4	28.0	49.1	
			c_4	10	1800.6	3804.2	2805.5	3072.0	25.9	

812

813 **Table 6.** Sample size (N), Minimum (Min), maximum (Max), mean, median and coefficient of variation
 814 (CV, in %) of the macroscopic capillary length, λ (mm), values.

Soil	Method	Type of data	Fitting method	Statistic					
				N	Min	Max	Mean	Median	CV
Sandy	Approach 1 and 3			186.9					
	Approach 2	Transient data	CI	8	0.2	28.7	10.2	5.1	114.6
			CL	6	0.8	14.2	6.8	6.5	81.1
			DL	7	0.9	188.7	47.0	8.6	153.9
	Steady data			$\lambda < 0$					
Approach 4			150.0						
Silty-loam	Approach 1 and 3			373.3					
	Approach 2	Transient data	CI	8	37.3	171.7	86.0	82.8	49.9
			CL	4	2.6	47.9	24.2	23.1	91.9
			DL	9	14.4	86.1	41.8	38.6	48.3
	Steady data			$\lambda < 0$					
Approach 4			150.0						
Sandy-loam 1	Approach 1 and 3			352.7					
	Approach 2	Transient data	CI	4	1.3	26.2	8.6	3.5	137.5
			CL	2	0.6	1.0	0.8	0.8	33.3
			DL	3	1.9	47.7	17.7	3.5	147.1
	Steady data			8	21.9	2410.9	741.1	265.2	131.4
Approach 4			150.0						
Sandy-loam 2	Approach 1 and 3			410.1					
	Approach 2	Transient data	CI	2	7.4	53.7	30.6	30.6	107.3
			CL	2	1.4	7.9	4.6	4.6	98.7
			DL	4	3.1	663.6	185.0	36.6	173.1
	Steady data			2	19.5	121.9	70.7	70.7	102.5
Approach 4			150.0						

815

816

817 **Table 7.** Sample size (*N*), Minimum (Min), maximum (Max), mean, median and coefficient of variation
818 (CV, in %) of the field-saturated soil hydraulic conductivity, K_{fs} (mm h⁻¹), values obtained by the SSBI
819 method, and the four approaches (1, 2, 3 and 4) for different data analysis procedures (transient and
820 steady data) and fitting methods (cumulative infiltration, CI, cumulative linearization, CL, and
821 differential linearization, DL).

Soil	Method	Type of data	Fitting method	Statistic					
				N	Min	Max	Mean	Median	CV
Sandy	Approach 1			10	52.8	127.6	91.0	91.2	24.9
	Approach 2	Transient data	CI	9	335.1	921.6	559.1	482.9	35.4
			CL	9	389.6	953.9	613.9	569.1	28.4
			DL	7	93.8	714.2	443.9	588.0	59.9
		Steady data		0					
	Approach 3	Transient data	CI	10	51.0	203.8	130.4	129.5	33.0
			CL	9	85.8	207.9	145.3	135.3	23.6
			DL	9	71.8	183.7	120.5	105.7	32.0
		Steady data		10	23.8	89.9	57.1	57.5	35.6
	Approach 4	Transient data	CI	10	60.2	240.8	154.1	153.1	33.0
			CL	9	101.4	245.7	171.7	160.0	23.6
			DL	9	84.9	217.1	142.4	125.0	32.0
		Steady data		10	28.1	106.2	67.5	67.9	35.6
	SSBI			10	32.3	121.9	77.4	77.9	35.6
Silty-loam	Approach 1			10	35.0	74.5	51.5	54.3	24.2
	Approach 2	Transient data	CI	8	130.1	252.2	186.0	188.7	21.5
			CL	10	175.2	971.4	600.2	663.6	43.7
			DL	9	180.1	517.6	302.9	275.7	41.0
		Steady data		0					
	Approach 3	Transient data	CI	10	23.7	77.6	53.7	53.1	29.1
			CL	10	41.5	118.1	78.1	84.2	31.1
			DL	10	33.1	90.6	60.4	62.9	29.8
		Steady data		10	24.4	50.0	34.9	36.6	24.9
	Approach 4	Transient data	CI	10	49.8	163.0	112.7	111.4	29.1
			CL	10	87.2	248.0	163.9	176.7	31.1
			DL	10	69.4	190.1	126.7	132.0	29.8
		Steady data		10	51.3	104.9	73.3	76.9	24.9
	SSBI			10	58.8	120.4	84.1	88.3	24.9
Sandy-loam 1	Approach 1			10	29.2	238.7	134.7	148.9	48.7
	Approach 2	Transient data	CI	10	495.7	5549.5	2321.5	2586.0	65.3
			CL	10	458.4	5938.9	2595.9	2774.2	59.0
			DL	8	673.3	3540.2	2089.7	2222.0	42.1
		Steady data		8	18.1	683.5	229.4	169.8	96.0
	Approach 3	Transient data	CI	10	43.8	448.9	212.9	247.5	56.6
			CL	10	41.6	476.7	225.2	249.4	55.7
			DL	9	54.2	310.7	207.0	221.1	39.7
		Steady data		10	23.2	166.7	95.6	106.8	46.1
	Approach 4	Transient data	CI	10	91.2	935.8	443.8	516.0	56.6
			CL	10	86.8	993.6	469.5	519.9	55.7
			DL	9	112.9	647.6	431.5	460.9	39.7
		Steady data		10	48.4	347.4	199.4	222.6	46.1
	SSBI			10	50.0	357.5	205.0	229.1	46.1
Sandy-loam 2	Approach 1			10	175.9	429.1	309.7	300.9	27.4
	Approach 2	Transient data	CI	7	1601.4	8352.6	5977.3	6538.4	38.9
			CL	8	3660.4	9230.1	6477.9	6569.4	29.3
			DL	7	284.3	7561.7	4308.6	4625.4	64.5
		Steady data		2	595.0	2156.7	1375.8	1375.8	80.3
	Approach 3	Transient data	CI	7	324.8	649.9	496.0	528.1	24.8
			CL	8	336.4	687.4	496.7	502.8	26.8
			DL	10	173.9	579.1	405.3	426.0	32.6
		Steady data		10	125.9	266.0	196.1	214.8	25.9
	Approach 4	Transient data	CI	7	776.7	1554.3	1186.1	1262.9	24.8
			CL	8	804.6	1643.8	1187.8	1202.4	26.8
			DL	10	415.9	1384.8	969.3	1018.7	32.6
		Steady data		10	301.0	636.0	469.1	513.6	25.9
	SSBI			10	309.8	654.5	482.7	528.5	25.9

823 **Table 8.** Sample size (N), minimum (Min), maximum (Max), mean, median and coefficient of variation (CV,
824 in %) of the unsaturated soil hydraulic conductivity, K_h (mm h^{-1}), values obtained at different pressure
825 heads, h_0 (mm) from the unit hydraulic gradient laboratory method for the four soils.

Soil	h_0	Statistic					
		N	Min	Max	Mean	Median	CV
Sandy	-10	3	130.1	176.3	154.7	157.8	15.0
	-30	5	85.7	150.9	115.1	106.5	21.6
	-60	5	41.2	115.6	88.1	96.3	31.9
	-120	5	8.6	58.3	39.4	46.9	48.4
Silty-loam	-10	5	11.6	29.2	21.0	24.0	37.1
	-30	5	8.4	21.7	14.0	12.4	38.2
	-75	5	2.9	11.7	7.4	7.7	43.2
Sandy-loam 1	-10	7	14.3	63.6	35.9	29.8	53.7
	-30	7	11.8	50.0	26.9	25.3	48.5
	-60	7	9.2	32.9	18.1	15.9	46.1
	-120	7	3.3	16.4	8.0	8.3	54.7
Sandy-loam 2	-10	9	38.5	183.2	87.1	65.4	62.9
	-30	9	22.5	104.9	50.4	48.0	52.7
	-60	9	12.0	64.2	29.9	30.5	55.7
	-120	9	4.8	21.9	11.5	10.4	45.4

826

827

828 **Figure 1. (a)** Procedure for estimating the equilibration time, t_s (h), and the infiltrated depth at the
829 equilibration time, $I(t_s)$ (mm), from cumulative infiltrations, and the intercept, c_3 (mm), and slope, c_4
830 (mm h^{-1}), of the regression line fitted to the last data points describing the steady-state conditions on the I
831 vs. t plot. **(b) (c) (d)** Estimation of the c_1 ($\text{mm h}^{-0.5}$) and c_2 (mm h^{-1}) parameters by the cumulative
832 infiltration (CI), cumulative linearization (CL) and differential linearization (DL) fitting methods. The
833 relative error, Er (%) [Eq. (15)], of the fitting of the functional relationships to the experimental data is
834 also reported. The example shows an infiltration run carried out at the silty-loam site.

835 **Figure 2.** Examples of the soil water content profiles at the final time of the experiments (t_{end}) and
836 infiltration curves modeled using the inverse solution from HYDRUS 2D/3D (dashed lines) compared
837 with the observed data (symbols) for the four soils. For each example, the water retention parameters h_b
838 (mm) and η (-), along with the field-saturated soil hydraulic conductivity, $K_{fs-HYDRUS}$ (mm h^{-1}), value
839 obtained by inverse solution from HYDRUS-2D/3D, and the root mean square error, $RMSE$ (mm), and
840 the relative error, Er (%), between the simulated and the observed curves are also reported.

841 **Figure 3.** Flowchart of the fourteen different scenarios and comparison between estimated and HYDRUS-
842 determined values (i.e., K_{fs} vs. $K_{fs-HYDRUS}$).

843 **Figure 4.** Illustrative examples showing three different abnormal behaviors of the infiltration curves.

844 **Figure 5.** Comparison between the mean c_1 ($\text{mm h}^{-0.5}$) and c_2 (mm h^{-1}) parameters estimated by the
845 cumulative infiltration (CI), cumulative linearization (CL) and differential linearization (DL) methods
846 for the four soils (sandy, silty-loam, sandy-loam 1 and sandy-loam 2). The relative error, Er (%), of the
847 fitting of the functional relationships to the experimental data is also reported. For a given variable and
848 soil, different letters represent significant differences according to the Tukey's Honestly Significant
849 Difference test ($P < 0.05$).

850 **Figure 6.** Cumulative empirical frequency distribution of the relative error of the field-saturated soil
851 hydraulic conductivity, $Er(K_{fs})$ [Eq. (14)], predictions (expressed as a percentage of the HYDRUS-
852 determined values, $K_{fs-HYDRUS}$) estimated by: i) Approach 1, ii) averaging individual determinations of the
853 four scenarios considered in the Approach 3, iii) Approach 4 with steady-state data analysis, and iv) the
854 SSBI method.

855 **Figure 7.** Individual value plot of differences between estimated and HYDRUS-determined values (i.e.,
856 $K_{fs} - K_{fs-HYDRUS}$). Gray and black circles indicate datasets that are respectively normally and non-normally
857 distributed according to the Kolmogorov-Smirnov test. Solid circles indicate datasets with a mean
858 (paired t -test) or median (Wilcoxon signed-rank test) difference between pairs not significantly different
859 from zero. Open circles indicate datasets with a mean or median difference between pairs significantly
860 different from zero.

861 **Figure 8.** Comparison between estimated and HYDRUS-determined values (K_{fs} vs. $K_{fs-HYDRUS}$).

862 **Figure 9.** Percentage of infiltration runs yielding a factor of difference, f_D , not exceeding 2, between 2 and 3,
863 and greater than 3, and percentage of failed runs. $f_D = MAX(K_{fs}, K_{fs-HYDRUS})/MIN(K_{fs}, K_{fs-HYDRUS})$.

864 **Figure 10.** Comparison for the four soils (sandy, silty-loam, sandy-loam 1 and sandy-loam 2) between the
865 mean unsaturated soil hydraulic conductivity, K_h (mm h^{-1}), values obtained at different pressure heads, h_0
866 (mm) from the unit hydraulic gradient laboratory method, and the mean field-saturated soil hydraulic
867 conductivity, K_{fs} (mm h^{-1}), values obtained by inverse solution from HYDRUS-2D/3D, the SSBI method,
868 and the four approaches (1, 2, 3 and 4), for different data analysis procedures (transient, Tr., and steady,
869 St.) and fitting methods (cumulative infiltration, CI, cumulative linearization, CL, and differential
870 linearization, DL). For each soil, the vertical list of scenarios reflect the descending order of the K
871 values.

872

















