



Achene traits involved in the water dispersal of the invasive *Fallopia x bohemica* complex: Variability and dispersal strategies

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► To cite this version:

Barbara Lamberti-Raverot, Florence Piola, Félix Vallier, Vanessa Gardette, Sara Puijalon. Achene traits involved in the water dispersal of the invasive *Fallopia x bohemica* complex: Variability and dispersal strategies. *Flora*, 2019, 251, pp.88-94. 10.1016/j.flora.2019.01.002 . hal-01988108

HAL Id: hal-01988108

<https://univ-lyon1.hal.science/hal-01988108>

Submitted on 1 Dec 2020

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1 **Achene traits involved in the water dispersal of the invasive *Fallopia* × *bohemica***
2 **complex: variability and dispersal strategies**

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14 **Keywords:** biological invasions, *Fallopia* × *bohemica*, floatability, germination, seed traits,

15 survival, water dispersal

Abstract

Secondary dispersal of terrestrial plant species through watercourses increases the dispersal range of floating seeds but their exposure to water may also challenge their viability and colonisation potential. Many terrestrial invasive species benefit from watercourses to extend their invasive range because water exposure can favour seed germination and early survival of seedlings for species sensitive to dry conditions and water flow enhances their dispersal distance downstream. Floatability, germination rate and survival after water exposure are seed traits involved in water dispersal and their variability may lead to different dispersal strategies. To address the question of the variability of achene traits involved in dispersal, we experimentally measured these three traits for *Fallopia × bohemica* achenes collected from 10 different stands. The floatability of achenes varied between stands from 2.5 to 5 days and their germination in water extended this period to 20 days through the emergence of floating seedlings. Germination proportions also differed between stands from 0.3 to 0.95 and were negatively related to the median germination time. Seedling survival after planting in soil was affected by their exposure time. This effect differed between stands, but a short germination time did not explain a better survival of some stands. These results show that achenes may differ in their dispersal potential in terms of floatability, as well as tolerance of seedlings to water exposure during dispersal. This study also highlights the fact that achenes of some stands present a combination of traits particularly adapted to long distance dispersal through watercourses, which can contribute to the extension of the invaded area.

1. Introduction

One of the factors influencing the spread of an invasive plant species through the landscape is its aptitude for dispersal (Theoharides and Dukes, 2007). A successful dispersal may be related to the propagule traits that increase the transport distance from the parental plant (Lloret et al., 2005; Theoharides and Dukes, 2007) and to the environmental characteristics of the recipient habitat. Many terrestrial invasive species benefit from watercourses to extend their invasive range (Catford and Jansson, 2014). For instance, water exposure can favour seed germination and early survival of seedlings for species sensitive to dry conditions (Funkenberg et al., 2012) and water flow enhances their dispersal distance downstream (Boedeltje et al., 2003). This is particularly true for buoyant propagules, such as seeds, which can disperse freely on the water surface, limiting their interaction with submerged obstacles, such as vegetation (Chang et al., 2008).

Buoyant seeds are also easily shed over higher zones of the riverbank (Coops and Velde, 1995). These zones are less subject to disturbances generated by floods and to continuous water exposure, which could challenge the colonisation of a terrestrial plant (Mahoney and Rood, 1998). The water dispersal of a terrestrial species may also positively affect seed germination and seedling establishment after deposition on the riverbank, as, for some species (e.g. *Populus* sp. and *Salix* sp.), this may accelerate seed imbibition and reduce the risk of drought stress on seedlings (Niiyama, 1990; Mahoney and Rood, 1998). However, water can also negatively impact viability, if seeds experience anoxic conditions or mechanical stress caused by water flow, which are frequently encountered by non-floating seeds (Casanova and Brock, 2000; Boedeltje et al., 2002). Young seedlings are generally more sensitive to water exposure, leading to reduced plant biomass that affects their survival (Niiyama, 1990; Casanova and Brock, 2000; Parolin, 2002).

The success of terrestrial species, in terms of dispersal distance and the effective number of dispersing seeds, may therefore depend on the interaction between the seed traits (e.g. floatability), survival and water exposure (Fumanal et al., 2007). Successful dispersal may be achieved using different strategies. For many riparian species, long floating periods that increase the dispersal distance (T_1 , Fig. 1A) are accompanied by delayed germination (T_2 , Fig. 1A), which diminishes the risks of seed loss if the seeds germinate during dispersal (S_1 , Fig. 1A, Clarke et al., 2001; van der Broek et al., 2005). For other species that present short floating periods, floating seedlings may extend dispersal time on the water surface (T_3 , Fig. 1B) by a rapid germination in water (Sculthorpe, 1967). In this case, the tolerance of seedlings to water exposure is critical (S_2 , Fig. 1B).

Previous studies have demonstrated that some species can produce two types of seeds presenting different dispersal strategies (e.g. short-floating vs. long-floating seeds, Pollux et al. 2009). This phenotypic variability constitutes an efficient mechanism for invasive species to adapt to broader environmental conditions and therefore extend their invaded range (Lee, 2002). Intraspecific variability in seed traits may therefore increase the dispersal potential in different habitats (Hantsch et al., 2013). For example, intraspecific seed heteromorphism affecting flotation and germination times could be involved in the dispersal of the terrestrial invasive species *Ambrosia artemisiifolia* through watercourses (Fumanal et al., 2007).

The *Fallopia* complex, one of the most invasive terrestrial taxa in the world (Lowe et al., 2000), presents a large phenotypic variability in achene morphology and flotation force (Lamberti-Raverot et al., 2017). Some studies on the role of sexual spread in the *Fallopia* complex have highlighted the potential role of watercourses in sexual dispersal, where this phenomenon is regularly present (Mandák et al., 2005; Bailey et al., 2009; Lamberti-Raverot et al., 2017). *F. × bohemica* produces a large number of fertile achenes each year (Engler et al., 2011; Groeneveld et al., 2014), which have the ability to remain afloat for 2-3 days and

germinating seedlings may also extend the floatation period (Engler et al., 2011; Rouifed et al., 2011). These results were seen only for one *Fallopia* stand, but given the large genetic variability of the complex (Bzdega et al., 2016), the variation in dispersal traits of *Fallopia* achenes between different stands and the combinations of traits leading to different dispersal strategies remain to be tested. To address this question, we experimentally measured the floatability, germination and survival rates after water exposure of *F. × bohemica* achenes and seedlings from 10 different stands. First, we tested if achene dispersal traits (floatation, i.e. T1 and T3, and germination timing, i.e. T2, Fig. 1) and propagule pressure (germination proportion and seedling survival after water exposure, i.e. S1 and S2, Fig. 1) differ between stands. Second, to assess the dispersal strategy of *Fallopia*, we tested if germination in water leads to an increased seedling dispersal time compared to that of the achenes ($T1 < T3$, Fig. 1), and whether seedling survival in soil is reduced due to extended water exposure ($S1 > S2$, Fig. 1). Finally, we tested if delayed germination (T2, Fig. 1) leads to the conservation of propagule pressure (S2, Fig. 1).

2. Materials and Methods

2.1 Species studied and sampling

The *Fallopia* complex includes *F. japonica*, *F. sachalinensis*, the complex of polyploid *F. × bohemica* hybrids (Chrtek & Chrtková) J.P. Bailey, originating from the hybridization of *F. japonica* and *F. sachalinensis*, backcrosses and F2, and other crosses involving *F. baldschuanica* (Bailey et al., 2009). Within the complex, *F. × bohemica* presents the highest genetic and morphological variation (Mandák et al., 2005; Tiébré et al., 2007). Although the spread of *F. × bohemica* has been mainly considered as clonal, a recent sexual spread has been shown in Europe (Engler et al., 2011; Rouifed et al., 2011, Buhk et al., 2015). Sexual

propagules consist of a single-seeded achene that is surrounded by a winged fruiting perianth. For the sake of simplicity, we use the term achene to designate the whole structure, i.e. the achene and the fruiting perianth.

The stands were identified as *F. × bohemica* and *F. japonica*. The achenes sampled were probably *F. × bohemica*, due to the male sterility of *F. japonica*, the large presence of *F. × bohemica* stands and the rarity of *F. sachalinensis* and *F. baldshuanica* in the region studied. Achenes were collected in November 2012 from 10 geographically distant stands, growing along seven different watercourses (Rhône-Alpes, France, Fig. 2). The sampling date was chosen relatively late in the season to maximize the chance that all the achenes were mature enough. All the stands were located on the river bank, very close to the water level. The stands sampled were characterized by high levels of achene production, except for two stands (BER and ELE) for which the achenes quantity at the sampling date was lower. As achene size may be relevant in dispersal (Lamberti-Raverot et al., 2017), sampling also reflected the range of achene sizes observed in the region. Achenes were randomly collected from several shoots in the same stand and stored under dry conditions at 4°C until they were used in the experiments.

2.2 Floatability and germination timing

Experiments started in February 2014. Floatability tests were performed under controlled conditions, in a climate room, to measure the floating time of achenes from the ten sampled stands. For each stand, 100 achenes were placed in two mesocosms (50 achenes per stand and per mesocosm). The mesocosms consisted of 35.5 × 26 × 7 cm (length × width × height) aquaria filled with tap water. Water was slightly agitated using perforated plastic tubing fixed around the bottom of the aquaria walls and connected to an air pump (0.6 L.h⁻¹). Water

agitation was used to allow good conservation of achenes throughout the experiment and prevent the achenes adhering to the aquarium walls and to themselves. The aquaria were illuminated using artificial light ($42 \mu\text{mol m}^{-2} \text{s}^{-1}$ surface light with a 12 h photoperiod) and the water temperature was maintained at $19.1 \pm 0.4 \text{ }^{\circ}\text{C}$. All the achenes were released simultaneously in the water, and the numbers of floating and sunken achenes were recorded twice a day for 11 days and then once a day until the end of the experiment (50 days). Germination of the achenes occurred in water, so the number of seedlings and their buoyancy (floating versus sunken) were also recorded.

2.3 Germination and survival after water exposure

To measure the effects of water exposure on seed germination and long-term seedling survival, 200 achenes per stand were placed in ten mesocosms (one per stand) placed in the same conditions as for measuring floatability and germination timing. Twenty achenes were collected at each of the following times: 0 (no exposure to water), 1, 2, 4, 8, 12, 20 and 28 days after achenes were placed in water. Immediately after collection, achenes were planted in four germination pots (five achenes per pot) filled with a mixture of potting compost (Favorit[®], NPK: 14.16.18) and sand (80/20). Germination pots were illuminated by artificial light ($42 \mu\text{mol m}^{-2} \text{s}^{-1}$ with a 12 h photoperiod) and were regularly watered throughout the experiment. After 4 to 12 days in water, most of the seeds had germinated so that the plants transferred to germination pots were mostly seedlings. The germination and survival rates of seedlings were recorded 3 times a week for 5 weeks.

2.4 Variable calculation and statistical analyses

After the first experiment, the count of floating achenes and seedlings, as well as the number of germinations, were used to estimate the following variables: median floatability of the achenes (T_1) and of the seedlings (T_3), as well as the median germination time in water (T_2) and the maximum germination proportion (G). These time-to-event data were described using four different four-parameter log-logistic models (drc R package, eq. 1):

$$f(x,(b,c,d,e)) = c + \frac{d - c}{1 + \exp\{b(\log(x) - \log(e))\}} \quad (\text{eq. 1})$$

We were specifically interested in three of the four parameters estimated by the model: b , c and e , where e represented the time producing a response half-way between the upper limit d and the lower limit c . T_1 was estimated as the parameter e of the first model. As all the floating seeds sank, the upper and lower limits of the model were fixed at 1 and 0, respectively. All ten stands were included in the model and T_1 was estimated for each stand. T_1 was compared between stands with the compParm function by means of ratios (the null hypothesis is that the ratio equals 1). T_2 and G were estimated as the parameters e and d of the second model, where the lower limit was fixed at 0. The ten stands were included in the model and T_1 and G estimated for each stand were both compared to each stand with the compParm function. T_3 was calculated as the difference between the estimated parameter e of seedlings floating again (third model) and the estimated parameter e of seedlings sinking (fourth model), which were independently estimated for each replicate ($n=2$) of each stand. Depending on the stand, either the upper or the upper and the lower limits of the models were fixed and the best model was chosen using the AIC criteria. The difference between stands was assessed by the Kruskal Wallis test as $n=2$. Student's t-test was used to compare both dispersal times, T_1 and T_3 .

After the second experiment, we tested the effect of water exposure on the survival proportion of seedlings for each stand using generalized linear models, with a binomial error. We then

estimated the probability of the seedlings surviving after planting in soil for two different cases of water exposure. In the first one, only achenes were exposed to water, corresponding to S_1 (Fig. 1): the probability of seedling survival after planting in soil was estimated for the longest time of achene exposure to water before they germinate. In the second one, both achenes and seedlings (after germination in water) could be exposed to water, corresponding to S_2 (Fig. 1): the probability of probability of seedling survival after planting in soil was estimated for the longest water exposure (i.e. 28 days, corresponding to the experimental duration). The probability of survival of S_1 and S_2 were compared to each other for each stand by performing χ^2 tests. A Pearson correlation test was carried out to test the relationship between T_2 and S_2 on the values estimated from the generalized linear models.

3. Results

The estimated median floatability of the achenes (T_1) differed significantly between the ten stands and varied from 2.5 to 5 days of flotation (Fig. 3). The median floatability of seedlings (T_3) was estimated for all the stands except one (BER) for which we could not fit a response curve. Seedling floatability for the nine other stands varied from 15 to 25 days and differences between stands were not significant ($\chi^2=13.2$, $df=8$, $p=0.1$, Fig. 3). Average seedling floatability was significantly longer than average achene floatability ($t=-8.86$, $df=8.49$, $p<0.001$, Fig. 3).

The two germination traits (estimated median germination times and germination proportion) differed between stands (Fig. 4). The proportion of germination in water (G) varied from 0.3 to 0.95 (Fig. 4A) and the median germination time (T_3) varied from 2 to 20 days (Fig. 4B). Regarding the survival traits, the survival probability of achenes germinating in soil after

water exposure (S_1) was almost 100% for the ten stands. Seedling survival was reduced after 28 days of water dispersal (S_2), but the effect of water exposure differed between stands (Fig. 5). The correlation between median germination time (T_2) and seedling survival after a long water exposure (S_2) was not significant ($t=0.86$, $df=8$, $p=0.4$).

4. Discussion

4.1 Trait variability

In accordance with our first hypothesis, some variability in achene traits involved in dispersal was observed. Variability of the *Fallopia* complex, and particularly of the genetically diverse *F. × bohemica* hybrid has been previously assessed for multiple sets of traits involved in growth (Pyšek et al., 2003; Herpigny et al., 2012), salt tolerance (Walls, 2010), secondary metabolite expression (Piola et al., 2013) or resistance to herbivory (Krebs et al., 2011). Although riparian ecosystems are some of the main natural dispersal vectors for *F. × bohemica* (Pyšek and Prach, 1993; Mandák et al., 2004; Bailey et al., 2007), this is the first study demonstrating the variability of achene traits that directly affect water dispersal.

Our analysis showed that most of the dispersal traits differed between stands, except median seedling floatability and seedling survival from achenes germinating in soil after water exposure. In the case of median seedling floatability, the floatability of seedlings ranged from 15 to 25 days and the lack of significant difference between stands may be partly due to a low statistical power. The morphological traits determining achene floatability (Carthey et al., 2016) seem to be more variable between stands than within, compared to those involved in seedling floatability. This difference could be related to variations in anatomical structure, related to the amount of air-filled tissues (Koch et al., 2010). Regarding S_1 , the lack of difference between stands suggests that for all the stands, the seeds equally protected the

embryos from water exposure. This is not surprising, as inhibition of germination has been shown to be a mechanism to protect the viability of dispersing seeds of many species, maximising the probability of seedling survival if germination occurs under favourable conditions (van den Broek et al., 2005).

The other four dispersal traits were highly variable. Achene floatability varied from 2.5 to 5 days between stands (Fig. 3). Differences in achene floatability may be related to different achene densities or surface tension (Carthey et al., 2016; Rendu et al., 2017). The proportion of achenes germinating in water differed between stands (30 to 95%), as well as the time to germination in water (from 2 to 20 days; Fig. 4). Seedling survival between stands only differed after long water exposure and ranged from 0.1 to 0.98 (Fig. 5). Larger seeds may increase seedling survival (Moles and Westoby, 2004; Skarpaas et al., 2011; Hantsch et al., 2013). However, in our study, germination-related traits and seedling survival after water exposure were not correlated to achene mass (data not shown). Instead, variations in seedling survival could be due to an intrinsic tolerance to low oxygen conditions of seedlings, as rapid starch degradation or enhanced ethanol fermentation (Ismail et al., 2009). Further research should focus on the determinants of seedling tolerance to water exposure and whether these mechanisms are related to the adaptation to *Fallopia* riverbank colonisation. Furthermore, the origin of the variability in achene germinability, traits and response to water exposure should also be investigated to determine the relative role of genetic background and environmental factors and phenology.

For the *Fallopia* complex, which exhibits a quick spread, hybridization may help to increase the genotypic diversity in the invasive area, leading to high phenotypic variations (Parepa et al. 2014), including those observed for seed traits. Phenotypic plasticity has been also observed in the vegetative traits of *Fallopia* (e.g. lamina size, growth and biomass allocation), which facilitate the colonisation of contrasting environments (Walls, 2010). In the case of

achenes, resource availability in the maternal environment may affect achene morphology and its subsequent dispersal and/or survival (Vaughton and Ramsey, 1998). Both mechanisms may therefore increase the variation of achene traits (Li and Feng, 2009; Hantsch et al., 2013), resulting in phenotypes with different water dispersal abilities. Large variations in dispersal traits may therefore favour the colonisation of different habitats, contributing to the invasiveness of the species along the rivers, but also, more generally, in terrestrial environments.

4.2 Dispersal strategies

Achene flotation times were relatively short (2.5 to 5 days) compared to the period after which 50% of the seeds had sunk (t_{50}) of 23 days measured by Boedeltje et al. (2003) for semi-aquatic and terrestrial plants. Our results are more consistent with the period after which 90% of the seeds had sunk (t_{90}) of 11.2 days measured for seeds defined as hydrochore or anemochore (Carthey et al., 2016). Achene flotation times among stands differed up to 100%. A 3-day difference in flotation may lead to greater dispersal distances in flowing water, and limits the risk of seed loss if environmental conditions do not lead to a rapid deposition on the riverbank (van den Broek et al., 2005). In addition, we have shown for all stands that the flotation time of achenes was significantly extended by germination in water, as expected in the second hypothesis and in accordance with Rouifed et al. (2011). Seedlings were able to remain afloat for over a month, compensating for short achene dispersal periods. This dispersal mechanism has been observed for several riparian species (Sculthorpe, 1967; Barrat-Segretain, 1996). *Fallopia* achenes germinate under water and seedlings develop cotyledons that increase the drag forces experienced (Koch et al., 2010), allowing them to be easily transported to the surface and benefiting from a secondary dispersal phase when temperatures start to rise.

In accordance with our second hypothesis, seedling survival was impacted by water exposure (Fig. 5). Before germination, seeds protect the embryo from mechanical damage and rotting due to the low-oxygen conditions of water to which seedlings are more exposed. A long germination period may participate in the conservation of achene viability during transport (van den Broek et al., 2005; Dacasa Rüdinger and Dounavi, 2008). Moreover, germination proportion is also negatively correlated to germination time, as previously demonstrated by Shipley and Parent (1991). If a long germination time helps to maintain survival levels, the propagule pressure of low germinating stands should be conserved after water exposure compared to high germinating stands. However, germination time and seedling survival were not significantly correlated as expected in our third hypothesis. This means that not only do achenes act to protect embryos from water exposure, but some mechanisms of seedling tolerance to water exposure may also support seedling survival. Moreover, seedling survival to long water exposure times differed greatly between stands. Indeed, half of the stands had a good tolerance to water exposure, and survival did not differ from that of achenes that germinated directly in soil. Tolerance to water exposure not only increases the dispersal potential, but also favours a rapid anchoring and colonisation at the deposition site (Scarano, 1998). For the other half, survival was significantly reduced, leading to low propagule pressure, despite the high intrinsic germination proportion of some stands. This means that water may act as an important selective force for individuals of terrestrial species, participating in long distance water dispersal.

Even if vegetative dispersal represents a significant part of the dispersal strategy of the *Fallopia* complex, sexual reproduction also has a major role, particularly in contributing to the increase in variability and therefore to the evolution of the taxon (Ellstrand and Schierenbeck, 2000; Parepa et al. 2014). Combining the measured achene traits did not lead to any contrasting dispersal strategies between stands, but some of the stands presented traits

that seem to be particularly adapted to dispersal along watercourses. The success of the dispersion also depends on the propagule pressure (i.e. the quantity and physiological condition of the propagules released, Lockwood et al., 2007), which increases the number of propagules released and effectively contributes to colonisation. Propagule pressure may differ between stands and also be affected by changes in propagule physiological conditions (germination, regeneration and survival aptitude), which alters the effective number of propagules able to establish (Lockwood et al., 2007). According to the spatial selection theory, better dispersers would benefit from reproductive isolation, leading to phenotypes exhibiting traits particularly adapted to long-distance dispersal (Shine et al., 2011), which was also observed for an invasive species colonisation process (Huang et al., 2015). This mechanism may contribute, in the case of *Fallopia*, to expanding the distribution range of the complex, through the evolution of its seed dispersal traits.

5. Conclusions

Our study demonstrated the variability of achene traits involved in dispersal and the differences in trait values observed between stands. This study also demonstrates that achenes of some stands present a combination of traits particularly adapted to long distance dispersal through watercourses, which can contribute to the extension of the invaded area. Sexual reproduction and hybridization may represent a permanent addition of highly invasive genotypes to the complex and contribute to the increase in variability and therefore to the evolution of the complex. The invasion potential of the complex is thus driven by a large phenotypic diversity leading to the emergence of new morphological or physiological traits. Large variations in dispersal traits and combination of traits may favour adaptation and

colonisation of different habitats, contributing to the invasiveness of the species along the rivers and in terrestrial environments.

Acknowledgements

We thank E. Jourdan, B. Oudot, F. Vallier and A. Vienney for their technical assistance. This work was supported by a grant from the Région Rhône-Alpes (N° 14 010108 01 ARC 2012 ADR) and by the CNR (Compagnie Nationale du Rhône). This study was carried out under the aegis of the Rhône Basin Long-Term Environmental Research (ZABR, Zone Atelier Bassin du Rhône). We thank two anonymous reviewers and the associate editor for their useful comments and suggestions.

References

- Bailey, J., Bímová, K., Mandák, B., 2009. Asexual spread versus sexual reproduction and evolution in Japanese knotweed *s.l.* sets the stage for the “Battle of the Clones”. *Biol. Invasions* 11, 1189-1203.
- Bailey, J.P., Bímová, K. and Mandák, B., 2007. The potential role of polyploidy and hybridisation in the further evolution of the highly invasive *Fallopia* taxa in Europe. *Ecol. Res.* 22, 920-928.
- Barrat-Segretain, M.-H., 1996. Strategies of reproduction, dispersion, and competition in river plants: A review. *Vegetatio* 123, 13-37.
- Boedeltje, G., Bakker, J.P., Bekker, R.M., Van Groenendael, J.M., Soesbergen, M., 2003. Plant dispersal in a lowland stream in relation to occurrence and three specific life-history traits of the species in the species pool. *J. Ecol.* 91, 855-866.

351 Boedeltje, G., Ter Heerdt, G.N., Bakker, J.P., 2002. Applying the seedling-emergence
352 method under waterlogged conditions to detect the seed bank of aquatic plants in
353 submerged sediments. *Aquat. Bot.* 72, 121-128.

354 Buhk, C., Thielsch, A., 2015. Hybridisation boosts the invasion of an alien species
355 complex: Insights into future invasiveness. *Perspect. Plant Ecol. Syst.* 17, 274–283.

356 Bzdega, K., Janiak, A., Ksiazczyk, T., Lewandowska, A., Gancarek, M., Sliwinska E.,
357 Tokarska-Guzik, B. A., 2016. Survey of genetic variation and genome evolution within
358 the Invasive *Fallopia* complex. *PloS one* 11(8), e0161854.

359 Carthey, A.J., Fryirs, K.A., Ralph, T.J., Bu, H., Leishman, M.R., 2016. How seed traits predict
360 floating times: a biophysical process model for hydrochorous seed transport behaviour
361 in fluvial systems. *Freshw. Biol.* 61, 19-31.

362 Casanova, M., Brock, M., 2000. How do depth, duration and frequency of flooding
363 influence the establishment of wetland plant communities? *Plant Ecol.* 147, 237-250.

364 Catford, J.A., Jansson, R., 2014. Drowned, buried and carried away: effects of plant traits
365 on the distribution of native and alien species in riparian ecosystems. *New Phytol.* 204,
366 19-36.

367 Chang, E.R., Veeneklaas, R.M., Buitenwerf, R., Bakker, J.P., Bouma, T.J., 2008. To move or
368 not to move: determinants of seed retention in a tidal marsh. *Funct. Ecol.* 22, 720-727.

369 Clarke, P.J., Kerrigan, R.A., Westphal, C.J., 2001. Dispersal potential and early growth in
370 14 tropical mangroves: do early life history traits correlate with patterns of adult
371 distribution? *J. Ecol.* 89(4), 648-659.

372 Coops, H., Velde, G., 1995. Seed dispersal, germination and seedling growth of six
373 helophyte species in relation to water-level zonation. *Freshw. Biol.* 34, 13-20.

374 Dacasa Rüdinger, M.C., Dounavi, A., 2008. Underwater germination potential of common
375 ash seed (*Fraxinus excelsior* L.) originating from flooded and non-flooded sites. Plant
376 Biol. 10, 382-387.

377 Ellstrand, N.C., Schierenbeck, K.A., 2000. Hybridization as a stimulus for the evolution of
378 invasiveness in plants? Proceed. Nat. Acad. Sci. 97, 7043-7050.

379 Engler, J., Abt, K., Buhk, C., 2011. Seed characteristics and germination limitations in the
380 highly invasive *Fallopia japonica* s.l. (Polygonaceae). Ecol. Res. 26, 555-562.

381 Fumanal, B., Chauvel, B., Sabatier, A., Bretagnolle, F., 2007. Variability and cryptic
382 heteromorphism of *Ambrosia artemisiifolia* seeds: What consequences for its invasion in
383 France? Ann. Bot. 100, 305-313.

384 Funkenberg, T. I. M., Roderus, D., Buhk, C., 2012. Effects of climatic factors on *Fallopia*
385 *japonica* s.l seedling establishment: evidence from laboratory experiments. Plant Species
386 Biol. 27(3), 218-225.

387 Groeneveld, E., Belzile, F., Lavoie, C., 2014. Sexual reproduction of Japanese knotweed
388 (*Fallopia japonica* s.l.) at its northern distribution limit: new evidence of the effect of
389 climate warming on an invasive species. Amer. J. Bot. 101, 459-466.

390 Hantsch, L., Bruelheide, H., Erfmeier, A., 2013. High phenotypic variation of seed traits,
391 germination characteristics and genetic diversity of an invasive annual weed. Seed Sci.
392 Res. 23, 27-40.

393 Herpigny, B., Dassonville, N., Ghysels, P., Mahy, G., Meerts, P., 2012. Variation of growth
394 and functional traits of invasive knotweeds (*Fallopia* spp.) in Belgium. Plant Ecol. 213,
395 419-430.

396 Huang, F., Peng, S., Chen, B., Liao, H., Huang, Q., Lin, Z., Liu, G., 2015. Rapid evolution of
397 dispersal-related traits during range expansion of an invasive vine *Mikania micrantha*.
398 Oikos 124, 1023-1030.

399 Ismail, A.M., Ella, E.S., Vergara, G.V., Mackill, D.J., 2009. Mechanisms associated with
 400 tolerance for flooding during germination and early seedling growth in rice (*Oryza*
 401 *sativa*). Ann. Bot. 103, 197–209.

402 Koch, E.W., Ailstock, M.S., Booth, D.M., Shafer, D.J., Magoun, A.D. 2010. The role of
 403 currents and waves in the dispersal of submersed angiosperm seeds and seedlings. Rest.
 404 Ecol. 18, 584-595.

405 Krebs, C., Gerber, E., Matthies, D., Schaffner, U., 2011. Herbivore resistance of invasive
 406 *Fallopia* species and their hybrids. Oecologia 167, 1041-1052.

407 Lamberti-Raverot, B., Piola, F., Thiébaud, M., Guillard, L., Vallier, F., Puijalon, S., 2017.
 408 Water dispersal of the invasive complex *Fallopia*: The role of achene morphology. Flora
 409 167, 150-157.

410 Lee, C.E., 2002. Evolutionary genetics of invasive species. Trends Ecol. Evol. 17, 386-391.

411 Li, Y.-P., Feng, Y.-L., 2009. Differences in seed morphometric and germination traits of
 412 crofton weed (*Eupatorium adenophorum*) from different elevations. Weed Sci. 57, 26-30.

413 Lloret, F., Médail, F., Brundu, G., Camarda, I., Moragues, E., Rita, J., Lambdon, P., Hulme,
 414 P.E., 2005. Species attributes and invasion success by alien plants on Mediterranean
 415 islands. J. Ecol. 93, 512-520.

416 Lockwood, J.L., Hoopes, M.F., Marchetti, M.P., 2007. Invasion ecology. Malden, US,
 417 Blackwell

418 Lowe, S., Browne, M., Boudjelas, S., De Poorter, M., 2000. 100 of the world's worst
 419 invasive alien species: a selection from the global invasive species database. Auckland,
 420 New Zealand, Invasive Species Specialist Group

421 Mahoney, J.M., Rood, S.B., 1998. Streamflow requirements for cottonwood seedling
 422 recruitment—an integrative model. Wetlands 18, 634-645.

423 Mandák, B., Bímová, K., Pyšek, P., Štěpánek, J., Plačková, I., 2005. Isoenzyme diversity in
424 *Reynoutria* (Polygonaceae) taxa: escape from sterility by hybridization. Plant Syst. Evol.
425 253, 219-230.

426 Mandák, B., Pyšek, P., Bímová, K., 2004. History of the invasion and distribution of
427 *Reynoutria* taxa in the Czech Republic: a hybrid spreading faster than its parents. Preslia
428 76, 15-64.

429 Moles, A.T., Westoby, M., 2004. Seedling survival and seed size: a synthesis of the
430 literature. J. Ecol. 92, 372-383.

431 Niiyama, K., 1990. The role of seed dispersal and seedling traits in colonization and
432 coexistence of *Salix* species in a seasonally flooded habitat. Ecol. Res. 5, 317-331.

433 Parepa, M., Fischer, M., Krebs, C., Bossdorf, O., 2014. Hybridization increases invasive
434 knotweed success. Evol. Appl. 7(3), 413-420.

435 Parolin, P., 2002. Submergence tolerance vs. escape from submergence: two strategies of
436 seedling establishment in Amazonian floodplains. Environ. Exp. Bot. 48, 177-186.

437 Piola, F., Bellvert, F., Meiffren, G., Rouifed, S., Walker, V., Comte, G., Bertrand, C., 2013.
438 Invasive *Fallopia x bohémica* interspecific hybrids display different patterns in
439 secondary metabolites. Ecoscience 20, 230-239.

440 Pollux, B.J.A, Verbruggen, E., Van Groenendael, J.M., Ouborg, N.J., 2009. Intraspecific
441 variation of seed floating ability in *Sparganium emersum* suggests a bimodal dispersal
442 strategy. Aquat. Bot. 90(2), 199-203.

443 Pyšek, P., Brock, J.H., Bímová, K., Mandák, B., Jarošík, V., Koukolíková, I., Pergl, J.,
444 Štěpánek, J., 2003. Vegetative regeneration in invasive *Reynoutria* (Polygonaceae) taxa:
445 the determinant of invasibility at the genotype level. Amer. J. Bot. 90, 1487-1495.

446 Pyšek, P., Prach, K., 1993. Plant invasions and the role of riparian habitats - A
447 comparison of 4 species alien to central Europe. J. Biogeogr. 20, 413-420.

448 Rendu, Q., Mignot, E., Riviere, N., Lamberti-Raverot, B., Puijalon, S., Piola, F., 2017.
449 Laboratory investigation of *Fallopia* × *bohemica* fruits dispersal by watercourses.
450 Environ. Fluid Mech. 17, 1051-1065.

451 Rouifed, S., Puijalon, S., Viricel, M.-R., Piola, F., 2011. Achene buoyancy and germinability
452 of the terrestrial invasive *Fallopia* × *bohemica* in aquatic environment: a new vector of
453 dispersion? Ecoscience 18, 79-84.

454 Scarano, F.R., 1998. A comparison of dispersal, germination and establishment of woody
455 plants subjected to distinct flooding regimes in Brazilian flood-prone forests and
456 estuarine vegetation. Oecol. Brasil. 4, 9.

457 Sculthorpe, C.D., 1967. Biology of aquatic vascular plants. Cambridge, UK, Edward
458 Arnold.

459 Shine, R., Brown, G.P., Phillips, B.L., 2011. An evolutionary process that assembles
460 phenotypes through space rather than through time. Proceed. Nat. Acad. Sci. 108, 5708-
461 5711.

462 Shipley, B., Parent, M., 1991. Germination responses of 64 wetland species in relation to
463 seed size, minimum time to reproduction and seedling relative growth rate. Funct. Ecol.
464 5, 111-118.

465 Skarpaas, O., Silverman, E.J., Jongejans, E., Shea, K., 2011. Are the best dispersers the best
466 colonizers? Seed mass, dispersal and establishment in *Carduus* thistles. Evol. Ecol. 25,
467 155-169.

468 Theoharides, K.A., Dukes, J.S., 2007. Plant invasion across space and time: factors
469 affecting nonindigenous species success during four stages of invasion. New Phytol. 176,
470 256-273.

471 Tiébré, M-S., Bizoux, J., Hardy, O., Bailey, J.P., Mahy, G., 2007. Hybridization and
472 morphogenetic variation in the invasive alien *Fallopia* (Polygonaceae) complex in
473 Belgium. Amer. J. Bot. 94(11), 1900-1910.

474 van den Broek, T., Diggelen, R., Bobbink, R., 2005. Variation in seed buoyancy of species
475 in wetland ecosystems with different flooding dynamics. J. Veget. Sci. 16, 579-586.

476 Vaughton, G., Ramsey, M., 1998. Sources and consequences of seed mass variation in
477 *Banksia marginata* (Proteaceae). J. Ecol. 86, 563-573.

478 Walls, R.L., 2010. Hybridization and plasticity contribute to divergence among coastal
479 and wetland populations of invasive hybrid Japanese knotweed sl (*Fallopia* spp.).
480 Estuaries and coasts 33, 902-918.

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483 **Figures**

484 **Figure 1.** Dispersal strategies observed in riparian ecosystems, in terms of time (T) during
485 which seeds disperse and germinate and the consequences for survival (S). A) Seeds may
486 show long floating periods (T_1) and long germination periods (T_2) that favour high seed
487 survival (S_1) or B) Short floating periods (T_1), rapid germination in water (T_2) and extended
488 dispersal time (T_3) through seedling flotation. In this case, the tolerance of seedlings to water
489 exposure is critical (S_2).

490 **Figure 2.** Study area and stands sampled. Each square represents a stand of *Fallopia* complex
491 sampled. The towns where the stands were located are Saint-Bernard (BER), Chavanay
492 (CHAV), Laveyron (ELE), Saint-Romain-en-Gier (ENT), Saint-Etienne-le-Molard (LIG),
493 Lozanne (LOZ), Montrond-les-Bains (MEY), Villieu-Loyes-Mollon (PCHA), Châtillon-la-
494 Palud (PGEV), Tencin (TEN).

495 **Figure 3.** Estimated median flotation values ($\pm$ SE) of achenes (T_1 , black bars) and seedlings
496 (T_3 , grey bars) of 10 stands of *F. x bohemica*. The hatched bar indicates the absence of the
497 median flotation time estimation for one stand. Different letters indicate significant
498 differences between estimated parameters ($p < 0.05$).

499 **Figure 4.** The right axis shows the estimated median germination times ($\pm$ SE, bars) and the
500 left axis the estimated germination proportion ($\pm$ SE, points) for 10 stands of *F. x bohemica*.
501 Different letters and numbers indicate significant differences between stands ($p < 0.05$).

502 **Figure 5.** Probability of seedling survival ($\pm$ SE) germinating in soil after water exposure (S_1 ,
503 grey bars) and after 28 days which was the maximum exposure duration (S_2 , black bars), for
504 10 stands of *F. x bohemica*. Different letters indicate significant differences between stands (χ

505 ² tests) and stars show significant differences (ns, not significant; **p < 0.01; ***p < 0.001)

506 between S₁ and S₂ for each stand.

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